

MICRO-CLIMATIC AMELIORATION IN A CALIFORNIA DESERT: ARTIFICIAL
SHELTER VERSUS SHRUB CANOPY.

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Abstract

Shrubs are structural agents of facilitation, providing benefit to animals that take refuge under their canopy. The central theme of this thesis is how shrubs compare to artificial shelters at reducing ambient temperature and incoming sunlight. We tested the effects of UV permeable artificial shelters on ambient temperature and solar radiation using two shapes (square and triangle) at three different blockage intensities (15%, 50%, and 90%), and contrasted those against the dominant shrub *Ephedra californica* and the open gap using temperature-light sensor loggers. Shelters offered more stable temperatures and shade from direct sunlight compared to the open gap and functioned analogous to *E. californica*. The square shelter best emulated the effects of *E. californica* on temperature and light. The findings of this thesis are useful for regional stakeholders as shelters can be used as a temporary refuge to increase thermal heterogeneity within the landscape whilst slow-growing, native vegetation is restored post-disturbance.

“After all, only the desert remains.”

- Zhu

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General Introduction

The study of interactions is the central theme of ecology. Competition, for the most part, dominated the field during the previous century; however, following the introduction of Bertness and Callaway's Stress Gradient Hypothesis (SGH), the focus has been shifted to facilitation as a fundamental interaction in many communities (Bertness and Callaway 1994; Bruno, Stachowicz, and Bertness 2003). Facilitation is defined as a positive interaction where one interacting species benefits, whilst none are harmed (Bertness and Leonard 1997). The SGH in particular proposes a shift from competition to facilitation with increasing stressful environmental conditions (Bertness and Callaway 1994). Facilitation is hence more common along a gradient of high abiotic stress as opposed to more benign conditions (Maestre et al. 2009). Due to this, many studies of positive interactions have focused on harsh environments, including arid ecosystems (Lu et al. 2018; Synodinos, Tietjen, and Jeltsch 2015; Maestre et al. 2009). Although the study of positive interactions is pivotal in harsh environments, it is almost paradoxical that non-trophic interactions are still ill-considered in predictive models of plant community responses to climate change (Anthelme, Cavieres, and Dangles 2014). Most studies of climate change focus on individual species, physiology, phenology, range shifts, and distribution, disregarding that an individual's responses to climate is connected through interactions with the same or neighbouring trophic levels (Walther 2010). Hence, to ideally advance the relative importance in stress, such as heat, aridity, and drought, with global change, we need to introduce non-trophic interactions into the equation when measuring climate and quantifying its impacts.

Foundational plant species are a vital component of facilitation research (Filazzola and Lortie 2014). Foundation species are locally abundant and common when compared to the keystone species (Attum and Eason 2006). Although they occupy lower trophic levels, similar to

keystone predators, they also create locally stable conditions required by other species. Foundational plants can include shrubs, nurse plants, perennials, trees, and cushion plants (Gómez-Aparicio et al. 2004). These vegetation have the ability to facilitate other taxa through mechanistic pathways that include, but are not limited to, seed trapping, abiotic stress amelioration, herbivore protection, increasing pollination services, facilitation-mediated secondary seed dispersal, and soil modification (Filazzola and Lortie 2014; Lortie, Filazzola, and Sotomayor 2016). Facilitation by shrubs is an established mechanism, able to repair and maintain semiarid ecosystems even post extensive damage (Lortie et al. 2018). Foundational plants have crucial impacts on the entire community dynamic; however, much of the research has focused solely on plant-plant interactions (Gómez-Aparicio et al. 2004; Castro et al. 2004; Flores and Jurado 2003), while the interaction with other taxa such as vertebrates is less-explored. There are a total of 63,000 documented species of vertebrates worldwide (Brown 2018). The south-western region of the United States is home to a variety of vertebrate species, including some of the first to be listed as endangered (Tazik and Martin 2002). Shrubs fulfill a critical role as agents of structural facilitation by offering an environment where animals can thermoregulate, reproduce, and take refuge (Lortie, Filazzola, and Sotomayor 2016; Filazzola et al. 2017). In order to advance the theory of facilitation, an emphasis also needs to be placed on examining direct and indirect shrub-animal interactions as well as shrub-plant, as they may be key to arid region management and restoration.

The state of California is home to a diverse array of vegetation including *Ephedra californica*. *Ephedra californica* is an ecologically dominant or co-dominant foundational shrub, wide-spread in hot deserts (Sawyer, Keeler-Wolf, and Evens 2009). In recent years however, climate change and extensive land-use has imposed severe stress on arid ecosystems, resulting in rapid degradation that may be difficult to reverse (Verwijmeren et al. 2013). In the south-western

region of the United States, anthropogenic disturbances and land-use have reduced the available terrestrial habitat, in turn decreasing biodiversity (Germano et al. 2011). The well-being and function of foundation plants species such as shrubs may depend on factors such as temperature, variability in precipitation, extended drought periods, and radiation (Tattini et al. 2006; Kogan and Guo 2015; MacDonald 2007). Given that landscape recovery post disturbance can be tremendously slow, it is simply unrealistic to merely rely on current management efforts to encourage the growth of new shrubs. Shrubs help augment structural diversity; thus, the availability of viable mimics as a temporary solution, alongside conservation efforts can enhance restoration outcomes.

Arid region expansion and desertification are important global change challenges (Asner and Heidebrecht 2005). Anthropogenic climate change significantly modifies physical and biological systems in all continents (Rosenzweig et al. 2008). Climate can be divided into micro and macro-climate. Micro-climate or weather can be defined as short-term (minutes to months) changes in atmospheric conditions in one small study site, while climate is the long-term weather pattern of a particular region (NASA 2005). By the year 2100, different micro-climatic parameters in California may vary; however, the overall temperature is predicted to increase by 5.6°-8.8°, which indirectly augments the frequency of extreme wildfires, and the average area burned statewide could increase by 77% (California's Fourth Climate Change Assessment 2019). Hence, to examine the ecological and biological relevance of climate, it is important to examine data at both levels because micro-climatic extremes can mediate day-to-day survival, but long-term patterns can impact reproduction and distribution (Bellard et al. 2012; Walther 2010). Climate can promote novel interactions between species (Parmesan and Yohe 2003). Climate envelope models are common tools to understand how species respond to change and environmental drivers, though one cannot ignore the interactions that buffer their tolerances.

The objective of this thesis is to examine the role shrubs as a form of structural facilitation and to test the efficacy of artificial shelters at mimicking the micro-climatic effects of foundation shrubs. In the following empirical chapter, I focus on the methodologies of UV Permeable Shade Cloth Shelters- a simple and cost-effective artificial canopy not examined previously. The goal of this chapter is not merely to describe how these shelters are built or how they're different from other prototypes discussed in the literature, but instead to understand the effects of shape and UV permeability on canopy micro-climate, including temperature and light intensity, relative to the open gap and the foundational shrub *E. californica*. I hypothesized that that artificial shelters can emulate the micro-climate of shrub canopies. The following predictions were tested:

- 1) The shape of the artificial canopy is an important characteristic that can influence abiotic parameters including temperature and solar radiation.
- 2) UV permeability influences the direct shade (light dappling) effects of a canopy on the thermal environment.
- 3) A combination of these factors can effectively emulate some of the effects of *E. californica*, provide cooler microsites, and decrease variation relative to the open gap areas.
- 4) Micro-climatic variation at finer scales (locally) differs from weather station data.

I confirm that shelters act similar to vegetation in order to increase the thermal heterogeneity within a given environment and are different from the paired, open gap microsites. I show that relationship between permeability/canopy structure and ambient temperature may be more complex than previously thought. Furthermore, I demonstrate that *in situ* logger data are ecologically as important to management as data recorded by weather-station. The concepts discussed in this thesis are valuable because globally no system is exempt from the impacts of

climate change. The key ideas discussed can be used as tools for stakeholders in various restoration strategies, in conjunction with other conservation and management practices.

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Abstract

Anthropogenic factors such as climate change, land use, urbanization, alongside the spread of invasive species are some of the challenges impacting the arid and semi-arid regions globally. The canopy of many native plants including shrubs and trees not only provides refuge from predators for some animals but also offers a shelter from climatic stressors for other plants. The canopy of native vegetation can thus be a microhabitat critical to the persistence of many species locally, and it is vital to better understand its importance for the conservation and recovery of species in these landscapes. In this study, we tested the hypothesis that triangular and rectangular artificial canopies function similarly to the canopy of resident native shrubs when ameliorating the understory micro-climate. Three light permeabilities including 15%, 50%, and 90% were tested by measuring soil and air temperature with light relative to paired open gap (non-canopied) microsites and shrubs. Shelters offered more stable temperatures and reduction in light compared to the open gap and were not significantly different from established native shrubs. This suggests that this simple, affordable intervention can provide a stop-gap solution that approximates natural heterogeneity in climate at fine scales and offers a refuge whilst managers and stakeholders restore native vegetation such as slow-growing and difficult to establish shrubs within this ecosystem.

Keywords: climate change, micro-climate, temperature, solar radiation, shelter, conservation, restoration.

Introduction

Climate change in arid and semi-arid regions is a critical issue globally. The rate of anthropogenic climate change is rapidly increasing in deserts and semi-arid grasslands (Williams 2014) and species need to adapt through many strategies. These changes in drylands in turn precipitate extensive ecological shifts including species loss (Barrows 2011), range shifts

(Bachelet et al. 2016), change in interactions (McCluney et al. 2012), increased invasion by exotic plants (Abatzoglou and Kolden 2011), and additional stress on resident species in these harsh environments (Finch 2012). Factors such as land-use changes including agriculture in drylands (Germano et al. 2011; Eliason and Allen 1997) can further decrease biodiversity by reducing the available terrestrial habitat for plants and animals (Nopper et al. 2018; Irwin et al. 2010; Elmqvist 2013). Furthermore, vegetation such as shrubs and other foundational plants are often removed or impacted (Sankey et al. 2012). In deserts, animals will not only experience large-scale changes such as drought, but also small-scale changes such as relatively more extreme fluctuations in abiotic factors such as temperature (Shrode and Gerking 1977; Hadley 1970). Deserts are getting hotter (Allen et al. 2014; Nabhan 2013) and long-term mega-droughts in some regions are relatively more frequent (Guerreiro, Kilsby, and Fowler 2017; Kogan and Guo 2015). This evidence suggests that not only do gross-scale changes in climate exert pressure on communities and sensitive species in drylands, but fine-scale changes and fluctuations can potentially further exacerbate local extirpation events (Olden, Poff, and Bestgen 2008), if not extinction (Munguia-Vega et al. 2013). Consequently, refuges, shelters, vegetation, or other attributes in the landscape are likely to enable persistence with changing climate through providing a buffer through variation by reducing the amplitude of variation.

Vegetation is a key aspect of most landscapes in drylands. Mechanistically, different types of vegetation are important for soil water retention as they could lead to different soil bulk densities in drylands (Wang et al. 2013). Shrubs are the dominant vegetation in deserts (Miriti, Joseph Wright, and Howe 2001; Throop et al. 2012). Shrub species can thus be used to examine climate change impacts and strategies used by associated plants and animals to adapt or respond to variation at fine-scales (Sotomayor and Drezner 2019). Foundation shrubs are able to facilitate

other taxa through various mechanistic pathways that include, but are not limited to, seed trapping, abiotic stress amelioration, herbivore protection, increasing pollination services, facilitation-mediated secondary seed dispersal, and soil modification (Filazzola and Lortie 2014; Lortie, Filazzola, and Sotomayor 2016). An important agent of structural facilitation is the shrub's canopy (Filazzola et al. 2017). Canopy micro-climates are generally cooler, more humid, and experience lower solar radiation compared to the open sites (Filazzola et al. 2017; Holzapfel and Mahall 1999). Shrubs fulfill a critical role; hence, more species are associated with shrubs than open spaces (Lortie, Filazzola, and Sotomayor 2016; Flores and Jurado 2003). Shrub canopies provide great benefits but can vary in extent that they filter or dapple light (Sonntag et al. 2007; Brantley and Young 2010), which likely influences the extent to which they cool the understory. Structural diversity alongside species diversity is crucial to many landscapes (Brooks 1999; Cowling et al. 1999; Morris 2000). Shrubs help increase structural diversity; hence, it is important to include them in the dialogue when discussing the impacts of anthropogenic changes on the landscape.

Heterogeneity in micro-climate and habitat at fine-scales is important of the maintenance of biodiversity. California is home to a diversity of dryland landscapes, dominated by many species of shrubs (Stuart and Sawyer 2001). *Ephedra Californica* (Mormon Tea) is a common foundation shrub species that benefits other plants (Lortie et al. 2018) and animals (Ivey et al. 2020). Shrubs provide a key component of this variation in many systems, including deserts (Fuhlendorf et al. 2017; Thorhallsdottir 1990). Hence, it is vital to advance the theory and application by A) testing shrubs as a thermal shelters and sources of fine-scale heterogeneity relative to open gap microsites, and B) directly test small, shrub-sized shelters both as a mean to more directly explore canopy effects without the biotic components of vegetation including litter, soil effects, or roots, and to examine a simple solution that can promote micro-heterogeneity in deserts and provide temporary

structural diversity. Furthermore, it is important to direct and sample the value of more shelters in some dryland systems as a form of thermal refuge and alternate modes of conservation whilst landscape recovery is made and new shrubs are grown.

Shrubs can be both keystone and foundation species in deserts. A keystone species (predator) is one that generally occurs in low abundances and occupies a high trophic level, but controls the density and diversity of other ecologically significant species (Mills and Doak 1993). On the other hand, a foundation species is locally abundant, common, and occupies a lower trophic level; yet, they also create locally stable conditions required by other species (Attum and Eason 2006). Foundation shrubs in dryland systems are typically slow-growing (Sawyer, Keeler-Wolf, and Evens 2009), difficult to establish in areas impacted by climate change (Meyer and Pendleton 2005), and frequently cleared by ranchers for livestock farming (Webb and Stielstra 1979). Hence, it would be ideal to have the capacity to mimic shrubs to augment and enhance low shrub cover areas and serve as stop-gap tools for conservation. Artificial canopies can provide an important surrogate test for canopy effects in drylands and there is a relatively long history of their use in ecology. Rainout shelters/drought nets and Open-Top-Chambers (OTC) have been used to study the change in a variety of abiotic parameters such as CO₂, temperature, soil temperature, solar radiation, and humidity (Yahdjian and Sala 2002; Marion et al. 1997). Although these shelters are effective, they can be expensive to build and may be difficult to assemble in a short period of time. Rainout shelters/drought nets used in semi-desert grassland studies have proven to be effective in altering precipitation, yet they have minimal impact on changing other variables such as air and soil temperature, humidity, and light (English et al. 2005; Gherardi and Sala 2013). On the other hand, OTCs have been experimentally used to increase temperature in plant studies in high-latitude ecosystems (Marion et al. 1997). Although these shelters are effective at manipulating different

abiotic parameters, they are typically larger in size (not shrub-sized). Additionally, large-scale solar farms can increase rain and vegetation (Li et al. 2018), but lead to habitat fragmentation thus limiting movement (Lovich and Ennen 2011). It is therefore key to take advantage of the variability in temperature and light in drylands to explore the effects of artificial shelters that are inexpensive and easily-built, which most importantly do not limit movement like solar farms or giant deploys, and help increase heterogeneity/open gaps that can be key to many animals.

Microsite heterogeneity through shrubs for climate is relatively novel concept for restoration and management, and provides an excellent framework to explore experimentally. Using a California desert ecosystem, we examined the hypothesis that artificial shelters can mimic the micro-climate of shrub canopies. The following predictions were tested: 1) Shelters are consistently cooler than open gap and not significantly different from shrub canopies, 2) shape of artificial shelters and UV permeability will shift light and temperature regimes- this is ecologically similar to light dappling effects of shrubs, 3) shelters have relatively lower variation in micro-climate in comparison to the open gaps, and 4) micro-climatic variation at these scales differs from similar sensors at weather stations. A deeper understanding of these physical structures impacts with and without other effects of living vegetation at fine-scales is important to better understanding habitat in deserts.

Materials & Methods

Study Site

This study was conducted in Panoche Hills Management Area located on the western edge of the San Joaquin Valley, California (Bureau of Land Management; 36°41.78' N, 120°47.89' W) (S; Supplementary Appendix). The regional climate can be characterized as arid/semi-arid. The average annual precipitation is 25.5 cm with an annual low and high temperature of 10.4 °C (50.72

°F) and 24.6 °C (76.3 °F), respectively (Filazzola et al. 2017). Winter and fall are considered to be the wettest seasons. The mean temperature observed in May is 20.4 °C (68.72 °F) and 23.7 °C (74.66 °F) in June (Los Baños Weather Station, <http://www.usclimatedata.com/>). The area is an *Ephedra californica* parkland, spread randomly between invasive grasses including *Bromus madritensis ssp. Rubens*, *Bromus hordeaceus*, *Erodium cicutarium* and *Schismus barbatus* (Filazzola et al. 2017). The study took place between May 20th to June 12th, 2019.

Microsite deployments

Shelters were constructed using PVC piping and UV permeable shade cloths at three permeabilities including 15%, 50%, and 90%. The open gap at 0% light blockage served as the procedural control as the frame was still deployed. The cloths were attached to the PVC using zip ties (Figure 1). Table A (Supplementary Appendix) describes the number of pieces at specific dimensions and diameter needed to build each triangle or square shelter. There were six replicates of each shape. Two for each blockage percentage for a total of 12 replicates. Pipes were slid onto metal stakes and secured into ground for stability (J and K; Supplementary Appendix). We selected a set of four microsites: shrub, open gap, square and triangle, and deployed all in the same area. Each microsite was geo-referenced (B; Supplementary Appendix). There were a total of 7 shrub-open pairs for a total of 14 microsites. Shrub canopy was measured at the x, y, and z plane where height (x) was the widest dimension of the canopy and perpendicular to the ground (Lortie et al. 2018). The open gap microsite was directly 2 meters away from the shrub at random orientation generated by a number table. The ground surrounding the shrub was mostly bare or contained patches of *Bromus madritensis ssp. Rubens*.

To measure the difference in light and temperature within canopied microsites and the open gap, Onset HOBO Temperature/Light Pendant (8K) loggers (Hoskin Scientific 2020) were placed

inside and directly outside to the right of the microsites. Each pendant was tied to a plastic stake using a zip tie, recording data at 1 hour intervals. Stakes were hammered into the ground until stable with ~10 cm remaining above ground. This was done to ensure that logger data were less-influenced by ground cover and true ambient conditions both inside and in the open were recorded. Air temperature (°F) and light intensity (lum/ft²) were recorded hourly. Loggers were placed out mid-May and collected in mid-June to account for spring-summer seasonal variation.

Shelters were constructed on-site. Rectangular (commonly referred to as square) shelters consisted of two sides with two 61 cm ½ inch pipes facing the ground connected to a 61 cm ¾ inch pipe using a 90° elbow. Triangular shelters were built using a 75 cm ¾ inch top pipe connected to a ½ inch to ¾ inch adapter. The adapter was then attached to a ½ inch 3-way 90° elbow fitted with two 61 cm ½ inch pipes. Shade Cloths were used to cover two sides of the triangular shelters and three sides of the rectangular shelters. The cardinal direction or orientation of each shelter was decided using a random number table and recorded. Shelters were inspected and repaired as needed throughout the study period.

Macro-climatic climate estimates

Hourly weather data were downloaded for the study site for the total duration of the study (Los Baños Weather Station at 37°03.30'N, 120°51.00'W, <http://www.usclimatedata.com/>). Date, air and soil temperature (°F) with solar radiation (W/m², converted to lum/ft²) were retrieved from this climate source, compiled, and published for re-use (Ghazian, Zuliani, and Lortie 2020).

Statistical analyses

All statistics were performed using R version 4.0.0 (R Core Team 2020). Code is published on Zenodo (Ghazian 2020) and micro-climate data are published on Figshare (Ghazian, Zuliani, and Lortie 2020). Q-Q plots were used to examine the distribution of data and to check for

normality and homoscedasticity (Schützenmeister, Jensen, and Piepho 2012). The relationship between temperature and light intensity was examined using Kendall’s rank correlation (non-parametric, continuous data). Generalized Linear Models (GLM) were used to compare temperature, light intensity, cover type, and microsite (Nelder and Wedderburn 1972). GLM dispersion parameters with AIC scores were used to compare and select the appropriate family to fit to models (Richards, Whittingham, and Stephens 2011). Temperature and solar radiation models were fit to Gaussian distribution. We explored spread in histograms by examining variance and used a Levene Test to check heterogeneity of variances for temperature and solar radiation across microsites (Schultz 1985). Post-hoc tests were done using the function *emmeans* from the *emmeans* R package (Lenth and Herve 2019). Relative Interaction Indices (RII) (Armas, Ordiales, and Pugnaire 2004) were used as an effect size measure to estimate the strength and direction of the microsite effect for temperature as follows:

$$RII = \frac{A_s - A_c}{A_s + A_c}$$

Where A_s and A_c are the parameters for ambient temperature under the shelter or shrub and the paired open gap microsite. The index values range from -1 to +1. For temperature, a positive value indicates that the shrub or shelter microsite is hotter relative to the open gap, or less ameliorated (Sotomayor and Drezner 2019). A value of 0 indicates a neutral effect.

Results

Temperature Effects

Shrub and the open gap microsites were consistently the warmest (Estimated Marginalized Mean (EMM) 73.9 ± 0.351 °F and 73.7 ± 0.219 °F, respectively), while the triangle and square microsites were the coolest (EMM 70.5 ± 0.467 °F and 72.7 ± 0.378 °F, respectively) (Figure 1, Table 2). The triangular shelter was the only microsite significantly cooler than the open gap (F;

Supplementary Appendix; post-hoc $p=0.0001$). Triangle was also significantly cooler than square (F; Supplementary Appendix; post-hoc $p=0.0034$). This cooling effect was most pronounced under the 90% blockage (I; Supplementary Appendix; post-hoc $p=0.0001$). Moreover, triangle was also significantly cooler than the shrub microsites (F; Supplementary Appendix; post-hoc test, $p=0.0001$). The square microsite had the lowest EMM for RII (-0.00308 ± 0.00408) (H, Q, and R; Supplementary Appendix). Higher maximum temperatures were more frequently recorded in the open gap microsites where relative variance was also the greatest (Figure 2, Table 3). We calculated the variance in temperature for each microsite, as well as for the weather-station data and found that they significantly differed (Figure 3, Table 3, Levene's F-Value= 60.096, $p=0.0001$). The lowest relative variance in temperature was observed at the weather station (Table 3), while the shrub had the highest variance of all canopied microsites (Table 3) followed by the open gap (Table 3). The lowest variance in temperature was seen under the square and triangle shelters (Table 3). Weather-station underestimated micro-climatic temperatures and were significantly cooler than the open gap, shrub, and square microsites (Figure 1 and Table 2, post-hoc test, $p=0.0001$) (F and L; Supplementary Appendix). Overall, temperature significantly increased with light intensity (Kendall's tau= 0.281, $p=0.0001$, N; Supplementary Appendix) and this relationship was significantly, positively linear at all microsites but not at triangle (O; Supplementary Appendix, $p=0.001$).

Light intensity effects

Daily mean light intensities were used to compare between microsites. The shrub microsite experienced the lowest light intensities (EMM 1628 ± 129.6 lum/ft²) followed by square (1675.5 ± 138.6 lum/ft²) and triangle (1861 ± 171.3 lum/ft²) (Table 2 and Figure 1). The open gap microsites experienced had the highest mean light levels (EMM 3331.5 ± 80.5 lum/ft²) (Table 2 and Figure

1). Square, triangle, and shrub experienced significantly lower light intensities compared to the open gap (Figure 1 and Table 2) (G; Supplementary Appendix; post-hoc test, $p=0.0001$). The light intensity under the square shelter was significantly lower than the triangle and the shrub (post-hoc $p=0.0001$; G; Supplementary Appendix). Furthermore, we looked at daily maximum solar radiation and found that the highest sunlight intensities were more often recorded by the weather-station and in the open gap (Figure 2). The relative variance in light experienced varied significantly between the microsites and the weather-station (Table 3, Levene's F-Value= 815.31, $p=0.0001$). The weather station experienced the highest variance in solar radiation (Table 3) followed by the open gap (Table 3), whilst the triangle, square, and shrub experienced lower variances (Table 3). Solar radiation measured at satellite weather-station was significantly higher than all microsites and the open gap areas (G; Supplementary Appendix; post-hoc test, $p=0.0001$).

Discussion

Shrubs and structural heterogeneity are important components of ecosystems relevant to the conservation and restoration of other plants and animals. A shelter, vegetation, or artificial of any sort in deserts provides amelioration or even just differences in the temperature and light at fine-scales that provides plants and particularly animals with thermal options (Ivey et al. 2020; Attum and Eason 2006). The hypothesis that artificial shelters can provide a similar thermal and light habitat to shrub canopies was supported here. Both shapes approximated the ameliorating canopy effects of the nearby *Ephedra californica*; however, square was most comparable to shrub at cooling as the two microsites did not statistically differ, and both were significantly hotter than triangle. The statistical difference in cooling between triangle and square was only significant under the 90% blockage. Shelter and shrubs significantly reduced the mean daily solar radiation relative to the open gap. Statistical spread for temperature was greatest in the open gap compared

to the shelter and shrub microsites. Daily temperature maxima were also the highest in the open gap. Sunlight experienced under the square canopy was significantly lower compared to shrub or triangle. Moreover, square and triangle experienced the lowest amplitude of variation. We also predicted that micro-climatic variation at microsite level differs from similar sensors at weather stations. Temperatures recorded at weather-station were significantly cooler than on-site level data. Additionally, solar radiation from weather-station was significantly higher than all other microsites (including the open gap) and the low spread experienced under shrubs or artificial canopies strongly suggested amelioration at fine-scales through buffered variation in climate. This evidence suggests that shelters can provide an important mechanism or tool for stakeholders to provide habitat for plants and animals either as a temporary stepping stone in restoration strategies or as a means to enhance habitat quality through simple and cost effective interventions.

Shrubs typically facilitate plants and animals within their understory. These canopies are able to ameliorate the physical conditions of the understory (Shumway 2000; Filazzola et al. 2017). These effects can influence annual density, species diversity (Kidron 2009) and associations with other taxa (Lortie, Filazzola, and Sotomayor 2016) including pollinators (Braun and Lortie 2019). Our shelters closely emulated the effects shrubs on understory by providing a cooler microhabitat with a lower daily maxima and amplitude of variation for solar radiation. This is consistent with findings of previous studies using artificial shades in drylands to create cooler and moister microhabitats (Smith, Patten, and Monson 1987; Barrow et al. 1996). Temperatures under the square canopy experienced the greatest amelioration effect and had the closest facilitation effect to that of *E. californica* perhaps because the canopy structure of a rectangular prism is more analogous to the canopy effects of *E. californica*. The structure of the canopy controls the quantity, quality, and temporal distribution of incoming sunlight, and that in turn impacts wind and air

movement, and subsequently temperature and precipitation regimes through boundary layer effects (Jennings, Brown, and Sheil 1998). Solar radiation is a direct thermospheric heating source and its effect on pressure differences results in solar wind fluctuations (Knipp, Tobiska, and Emery 2004). In natural vegetation, leaf area index (LAI) is a dimension-less value of the leaf area per unit ground area (Breda 2003). The cooling effect on canopy air temperature and shade effectiveness is directly related to LAI, with species of higher LAI values generally providing a greater cooling effect compared to other species (Tukiran 2016). However, the relationship between LAI and canopy cover is not clear-cut and can be more complex than previously thought (Nielsen, Miceli-Garcia, and Lyon 2012). In non-deciduous plants, LAI may be more related to branching and twigs, rather than the leaf itself, with thicker branching resulting in higher LAIs; therefore, influencing the amount of incoming sunlight (Wilfong, Brown, and Blaser 1967). This can also change light quality in the ratio of red to far red/blue relevant to other plant species (Kasperbauer 1971) and generate light dappling effect that can influence plants and animals through relatively lower and more variable intensities (Brantley and Young 2010), further influencing direct cooling effects. Similar to vegetation, we suspect that the geometrical structure of a rectangular prism (canopy volume and depth) and sunflecking through shade clothes at 15% and 50% most similarly matches the canopy effects and possibly the LAI of *E. californica* (though, the latter cannot be confirmed since LAI was not measured). The combined effects of shape and shade clothes influence wind movements and increase indirect cooling effect. This suggests that artificial canopies can in fact be used in conjunction with shrubs or in shrub-less areas in order to enhance ecosystem well-being in times of high abiotic stress.

The scale at which you measure climate is important for plants and animals. We found that mean daily temperatures recorded at the nearby weather-station were significantly lower than

microsite level logger data and had less variation compared to shrub and shelter microsite. Moreover, daily maximum temperature recorded by weather station were always lower than all fine-scale level data. A study by Lathlean et al. (2011) reported significantly lower air temperatures when measured via *in situ* loggers and concluded that coarse-scale data were ineffective at capturing extremes in air temperature variability. Additionally, the ground at Panoche Hills is not completely bare and is, in fact, covered with a thick, dry layer of golden-coloured Mediterranean grasses during the spring and summer periods that reflect light back to the loggers; therefore, increasing the on-site recorded temperature. Desertification from sustained drought periods can decrease vegetation greenness and make surfaces look lighter, lowering the shadowing effect and hence increasing land surface albedo (Ghulam et al. 2007). This is consistent with the idea that dark surfaces typically absorb more incident radiation than light-coloured, high reflectance surfaces (Stuart-Fox, Newton, and Clusella-Trullas 2017). This discrepancy between weather-station and on-site level data goes to demonstrate that locally climate is experienced differently compared to satellite-level data. Hence, it is ecologically vital that managers not only considers coarse-scale climate in conservation practices, but also incorporate micro-climate since only considering station data may be detrimental to the survival of small plants and animals that are more susceptible to temperature fluctuations and thermal exasperation.

Implications

Micro-environmental variation is a form of habitat and thermal and structural heterogeneity is critical for some animals including ectotherms, in addition to providing refuge (Bauwens, Hertz, and Castilla 1996; Diaz and Cabezas-Diaz 2004). There is an important positive relationship between species richness and environmental heterogeneity, and environmental heterogeneity can also impact community dynamics (Yang et al. 2015). Environmental heterogeneity can increase

the potential species of species being able to colonize different microsites (Lundholm 2009). In a way, micro-environmental heterogeneity provides different niches for a variety of species (Kadmon and Allouche 2007) or resources for habitat selection processes (Lortie et al. 2020; Boyce and McDonald 1999). Shelters can function like vegetation in some capacity and thus increase the thermal heterogeneity within a given environment, at least in deserts. In California, climate change is interfering with wildfire regimes and altering biological communities (Bishop et al. 2019). Not only can post-disturbance recovery of vegetation take decades (Berry et al. 2016), but competition and invasion by non-natives are amongst other challenges slowing the recruitment of native vegetation (Bowman et al. 2009, 2011). Hence, the benefit of artificial shelters as a mode of conservation is evident, whilst other efforts are made to re-establish the native community and the natural vegetation has had the time and resources to re-emerge. Signs of human-induced climate change is already visible in a variety of ecosystems. Species all around the world face changes in distribution and abundance due to migration and range shift (Midgley et al. 2002). This change will impact the physiology, growth, and productivity of biota (Cannell 1998), as well as their behaviour (Walther, Burga, and Edwards 2001). Given the current rates, it will not be long before species can no longer physiologically and behaviourally mitigate the impacts of climate change. Animals such as lizards may already be over-expending energy when trying to thermoregulate (Vickers, Manicom, and Schwarzkopf 2011). This study suggests that shelters offer a mechanism to create climate refuges as a temporary solution or a long-term strategy, and as an effective form of interference for today's every-growing anthropogenic disturbances.

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Figures & Tables

Table 1. Key contrast of abiotic measurements estimated using GLM for the study period. (May 20th to June 12th, 2019). Microsite/site and cover type were each treated as a factor. Significant P-values are in bold. See Methods for model fitting.

		<i>df</i>	Deviance Resid.	<i>df</i> Resid.	Dev	Pr(>Chi)
<u>Measure:</u>						
Temperature	NULL			22583	10784021	
	as.factor (microsite)	5	74212	22578	10709809	0.0001
Solar Radiation/Inte nsity	NULL			13253	6.42x10 ¹¹	
	as.factor (microsite)	5	1.33x10 ¹¹	13248	5.09x10 ¹¹	0.0001
Temperature by Blockage (cover type %)	NULL			15388	7168340	
	as.factor (microsite)	2	18912.1	15386	7149427	0.0001
	as.factor (cover type)	3	16739.0	15383	7132688	0.0001
	as.factor (microsite): (cover.type)	2	2370.6	15381	7130318	0.07755
Relative Interaction Index (RII)	NULL			61	0.025029	
	as.factor (microsite)	2	0.0014171	59	0.023612	0.1703

Table 2. Estimated Marginalized Mean (EMM) and standard error (SE) are given for each microsite and weather station based on temperature (°F) and solar radiation (lum/ft²) GLM. Confidence Interval used is 95%.

<i>Measurement</i>	<i>Microsite/Site</i>	<i>emmean</i>	<i>SE</i>	<i>Asymp.LCL</i>	<i>Asymp.UCL</i>
Temperature (°F)	open	73.7	±0.219	73.3	74.1
	shrub	73.9	±0.351	73.2	74.6
	shrub.surface	77.0	±0.417	76.2	77.8
	square	72.7	±0.378	71.9	73.4
	soil.surface				
	triangle	70.5	±0.463	69.6	71.4
	weather.station	68.3	±0.872	66.6	70.0
Solar Radiation (lum/ft²)	open	3331.5	±80.5	3174	3489
	shrub	1628	±129.6	1374	1882
	soil.surface	33.9	±186.8	-332	400
	square	1675.5	±138.6	1404	1947
	triangle	1861.4	±171.3	1526	2197
	weather.station	16478	±248.2	15992	16964

Table 3. Spread calculated using variance in temperature and solar radiation at microsities and site. Significant P-values are bolded.

<i>Microsite</i>	<i>Variance Temperature (s²)</i>	<i>Variance Solar Radiation (s²)</i>
open	786.0095	51449536.56
shrub	1022.394	11887055.49
soil.surface	955.08566	23770.813
square	698.6578	10107115.06
triangle	488.1092	28630191.11
weather.station	212.2233	5.38x10 ⁸

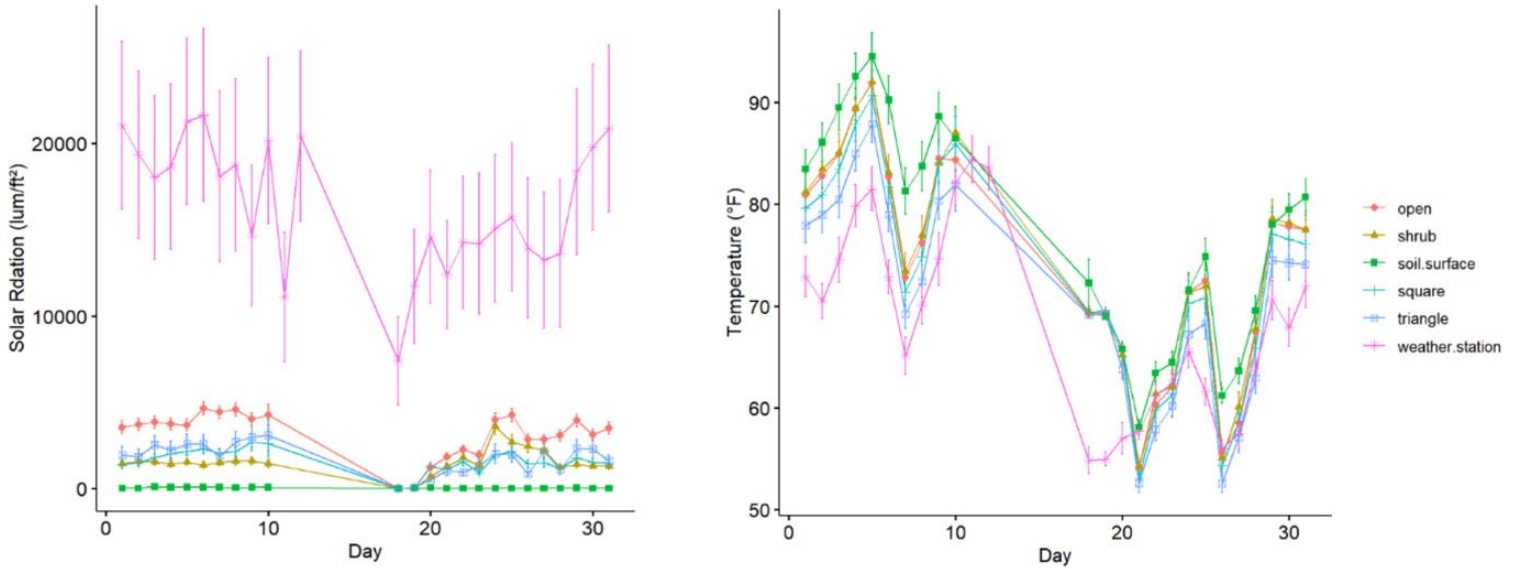


Figure 1. Mean daily temperature (°F) and solar radiation (lum/ft²) over the course of 2019 spring-summer season recorded at each microsite using micro-loggers and retrieved from *Los Baños* Weather Station. Soil temperature and solar radiation were recorded using loggers. Point shapes represent different microsites. Solid lines connect daily means. Errors bars are standard error (SE).

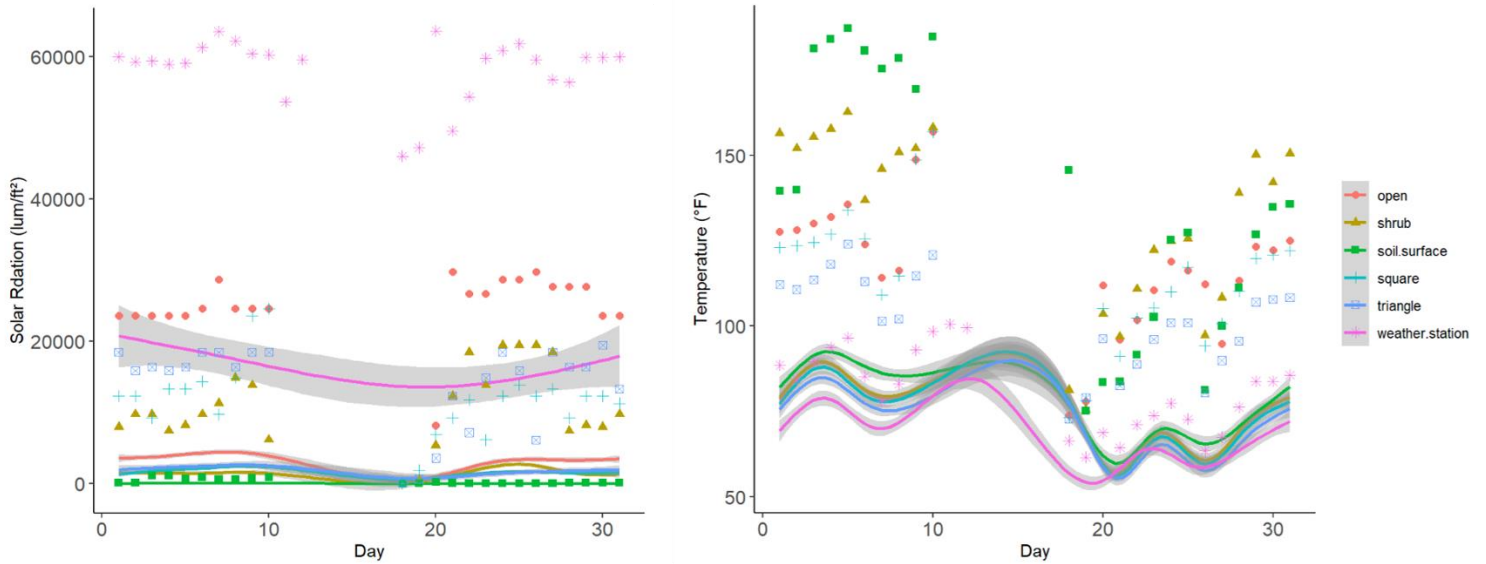


Figure 2. Smoothed conditional mean for temperature (°F) and solar radiation (lum/ft²) over the course of the spring-summer 2019 season recorded at each microsite and retrieved from *Los Baños* Weather Station. Soil temperature and solar radiation were recorded using loggers. Point shapes represent maximum dailies at each microsite. Confidence interval shown are standard error.

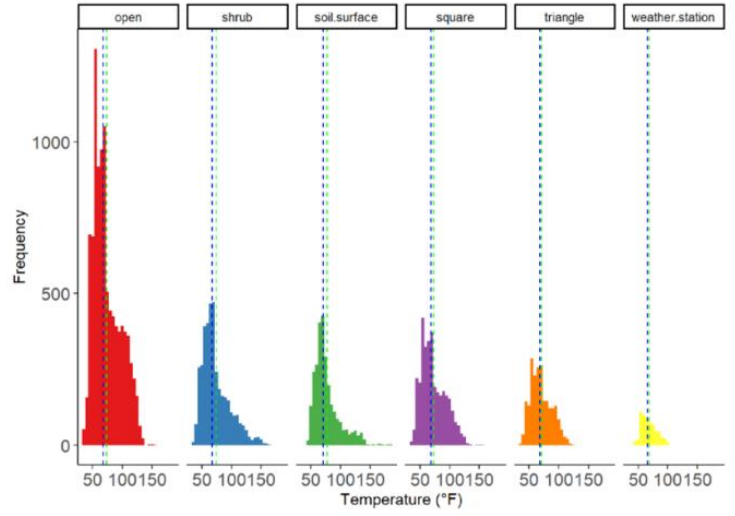
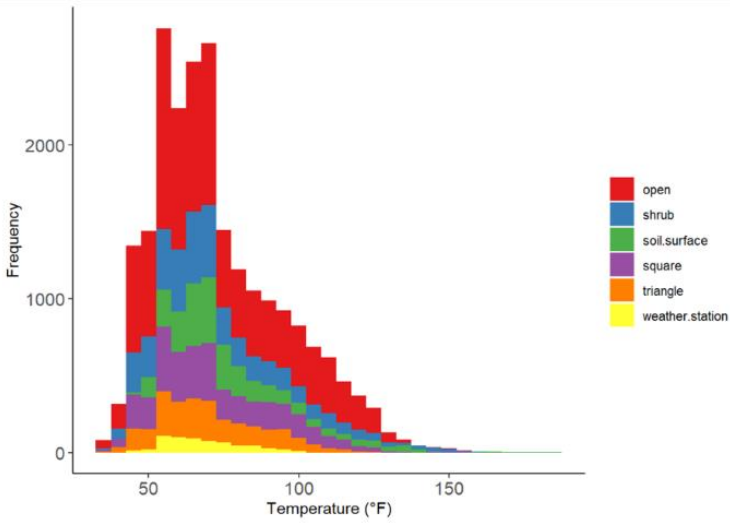


Figure 3. Frequency histogram of temperatures (°F) recorded at each microsite or weather station and combined. Vertical green dashed lines are the mean and the blue dashed lines represent the median. Soil temperature was recorded using loggers.

Supplementary Appendix

A. List of PVC pieces used for shelter skeleton construction is provided alongside the quantity needed to build one of each shelter-type.

<i>Piece</i>	<i>Quantity for Triangular Shelter</i>	<i>Quantity for Rectangular Shelter</i>
61 cm (½ inch diameter) pipe	4	4
61 cm (¾ inch diameter) pipe	NA	2
75 ¾ cm pipe	1	NA
½ inch to ¾ inch adapter	2	NA
½ inch to ¾ inch 2-way 90° elbow	NA	4
½ inch 3-way 90° elbow	2	NA

B. Location (latitude and longitude coordinates) of each shelter-open and shrub-open gap microsite is given, alongside its shape and cover type (if applicable).

<i>Shelter ID</i>	<i>Latitude</i>	<i>Longitude</i>	<i>Shape (Triangle/Square)</i>	<i>Cover type</i>
1	36.69363	-120.79318	T	15%
2	36.69364	-120.79331	S	15%
3	36.69355	-120.79315	S	90%
4	36.69349	-120.79320	T	90%
5	36.69349	-120.79311	T	50%
6	36.39342	-120.79311	S	50%
7	36.69394	-120.79300	S	15%
8	36.69397	-120.79292	T	15%
9	36.69401	-120.79282	S	90%
10	36.694	-120.79295	T	90%
11	36.69405	-120.79305	S	50%
12	36.69408	-120.79301	T	50%
<i>Shrub ID</i>				
1	36.69532	-120.797		
2	36.69592	-120.797		
3	36.69533	-120.794		
4	36.69598	-120.797		
5	36.69591	-120.797		
6	36.69605	-120.797		
7	36.69595	-120.798		

C. Generalized Linear Model (GLM) summary for temperature (°F) at each microsite/site. 95% Confidence Intervals are provided along with the P-value for each microsite. Significant P-values are bolded.

<i>Predictors</i>	<i>Estimates</i>	temp	
		<i>CI</i>	<i>p</i>
(Intercept)	73.69	73.26 – 74.12	<0.001
microsite [shrub]	0.25	-0.56 – 1.06	0.551
microsite [soil.surface]	3.34	2.41 – 4.26	<0.001
microsite [square]	-1.00	-1.86 – -0.15	0.022
microsite [triangle]	-3.19	-4.20 – -2.19	<0.001
microsite [weather.station]	-5.42	-7.18 – -3.66	<0.001
Observations	22584		
R ² Nagelkerke	0.963		

D. Generalized Linear Model (GLM) for solar radiation (lum/ft²). 95% Confidence Intervals are provided along with the P-value for each microsite. Significant P-values are bolded.

<i>Predictors</i>	<i>Estimates</i>	intensity	
		<i>CI</i>	<i>p</i>
(Intercept)	3331.50	3173.71 – 3489.29	<0.001
microsite [shrub]	-1703.47	-2002.52 – -1404.43	<0.001
microsite [soil.surface]	-3297.64	-3696.23 – -2899.05	<0.001
microsite [square]	-1656.02	-1970.22 – -1341.83	<0.001
microsite [triangle]	-1470.12	-1841.06 – -1099.17	<0.001
microsite [weather.station]	13146.49	12635.12 – 13657.86	<0.001
Observations	13254		
R ² Nagelkerke	1.000		

E. Generalized Linear Model (GLM) for the relationship between temperature (°F) and solar radiation (lum/ft²) by microsite/site. 95% Confidence Intervals are provided along with the P-value for each microsite. Significant P-values are bolded.

<i>Predictors</i>	<i>Estimates</i>	temp	
		<i>CI</i>	<i>p</i>
(Intercept)	79.43	78.84 – 80.02	<0.001
intensity	0.00	0.00 – 0.00	<0.001
microsite [shrub]	1.00	-0.12 – 2.12	0.079
microsite [soil.surface]	9.63	8.27 – 10.99	<0.001
microsite [square]	-1.88	-3.08 – -0.68	0.002
microsite [triangle]	-3.48	-4.82 – -2.15	<0.001
microsite [weather.station]	-16.17	-18.21 – -14.12	<0.001
intensity * microsite [shrub]	0.00	0.00 – 0.00	<0.001
intensity * microsite [soil.surface]	0.10	0.09 – 0.11	<0.001
intensity * microsite [square]	0.00	0.00 – 0.00	<0.001
intensity * microsite [triangle]	0.00	-0.00 – 0.00	0.768
intensity * microsite [weather.station]	-0.00	-0.00 – -0.00	<0.001
Observations	13254		
R ² Nagelkerke	1.000		

F. Post-hoc analysis of microsites based on temperature GLM. Standard error (SE) and P-values are given. Significant P-values are bolded.

<i>Contrast</i>	<i>estimate</i>	<i>SE</i>	<i>z.ratio</i>	<i>P-value</i>
open-shrub	-0.247	0.414	-0.597	0.9913
open-soil.surface	-3.337	0.471	-7.077	0.0001
open-square	1.005	0.437	2.298	0.1947
open-triangle	3.195	0.512	6.238	0.0001
open-weather.station	5.42	0.899	6.029	0.0001
shrub-soil.surface	-3.09	0.545	-5.665	0.0001
shrub-square	1.252	0.516	2.426	0.1473
shrub-triangle	3.442	0.581	5.952	0.0001
shrub-weather.station	5.667	0.94	6.030	0.0001
soil.surface-square	4.342	0.563	7.708	0.0001
soil.surface-triangle	6.532	0.623	10.480	0.0001
soil.surface-weather.station	8.757	0.967	9.059	0.0001
square-triangle	2.190	0.598	4.646	0.0034
square-weather.station	4.415	0.95	4.646	0.0001
triangle-weather.station	2.225	0.987	2.254	0.2131

G. Post-hoc analysis of microsites based on solar radiation GLM. Standard error and P-values are given. Significant P-values are bolded and confidence level used is 95%.

<i>Contrast</i>	<i>estimate</i>	<i>SE</i>	<i>z.ratio</i>	<i>P-value</i>
open-shrub	1703.5	153	11.165	0.0001
open-soil.surface	3297.6	203	16.215	0.0001
open-square	1656	160	10.330	0.0001
open-triangle	1470.1	189	7.768	0.0001
open-weather.station	-13146.5	261	-50.387	0.0001
shrub-soil.surface	1594.2	227	7.013	0.0001
shrub-square	-47.5	190	-0.25	0.9999
shrub-triangle	-233.4	215	-1.086	0.8870
shrub-weather.station	-14850	280	-53.039	0.0001
soil.surface-square	-1641.6	233	-7.058	0.0001
soil.surface-triangle	-1827.5	253	-7.212	0.0001
soil.surface-weather.station	-16444.1	311	-52.944	0.0001
square-triangle	-185.9	220	-0.844	0.9593
square-weather.station	-14802.5	284	-52.072	0.0001
triangle-weather.station	-14616.6	302	-48.472	0.0001

H. Estimated Marginalized Mean (EMM) and standard error (SE) are given for each microsite based on RII GLM. Results. Confidence Interval used is 95%.

<i>Microsite</i>	<i>emmean</i>	<i>SE</i>	<i>Asymp.LCL</i>	<i>Asymp.UCL</i>
shrub	0.00309	±0.00535	-0.007393	0.01357
square	-0.00308	±0.00408	-0.011088	0.00492
triangle	0.00776	±0.00408	-0.000245	0.01576

I. *Emmeans* pairwise contrast of ambient temperature at different microsites by cover type (blockage %). Results are given at 95% Confidence Interval. Bolded P-value(s) means significantly different.

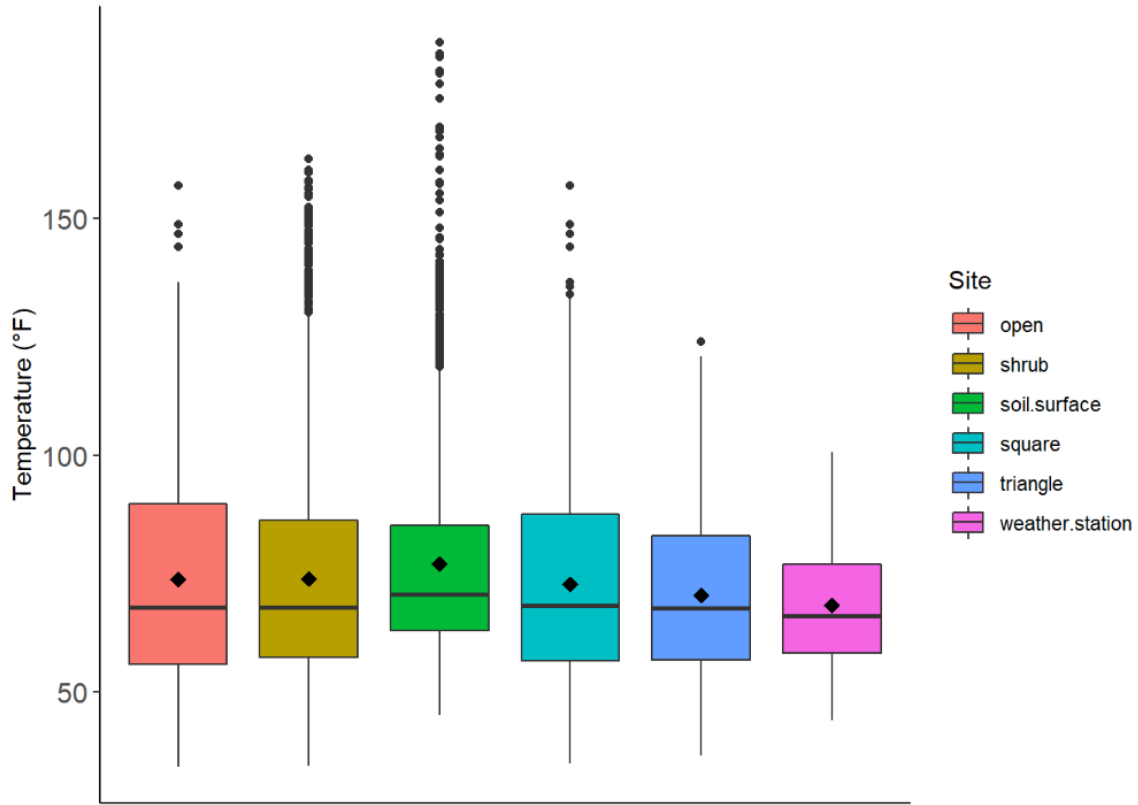
<i>Cover Type</i>	<i>Contrast</i>	<i>estimate</i>	<i>SE</i>	<i>z.ratio</i>	<i>p-Value</i>
0	Open-triangle	-26.593	±21.532	-1.235	0.4324
15	Square-triangle	1.031	±1.126	0.916	0.6301
50	Square-triangle	0.584	±1.123	0.52	0.8616
90	Square-triangle	3.527	±3.527	3.853	0.0003



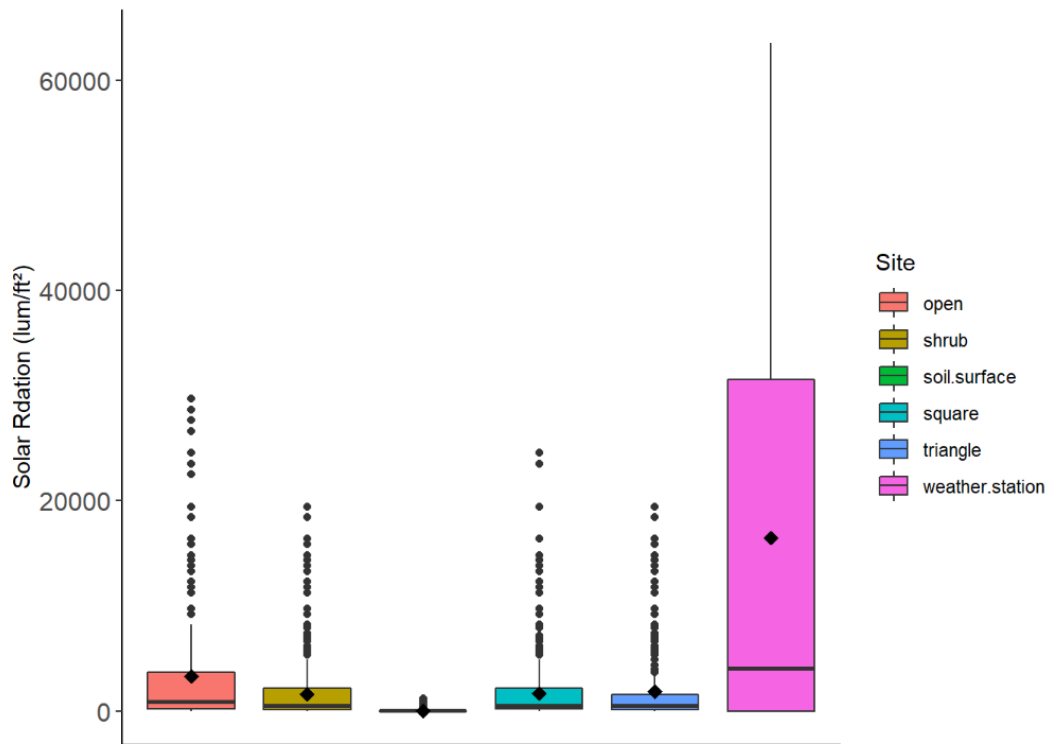
J. Left-Triangular shelter with 90% shade cloth attached to PVC skeleton using zip ties. Right-Rectangular shelter with 15% shade cloth attached to two PVC skeletal frames.



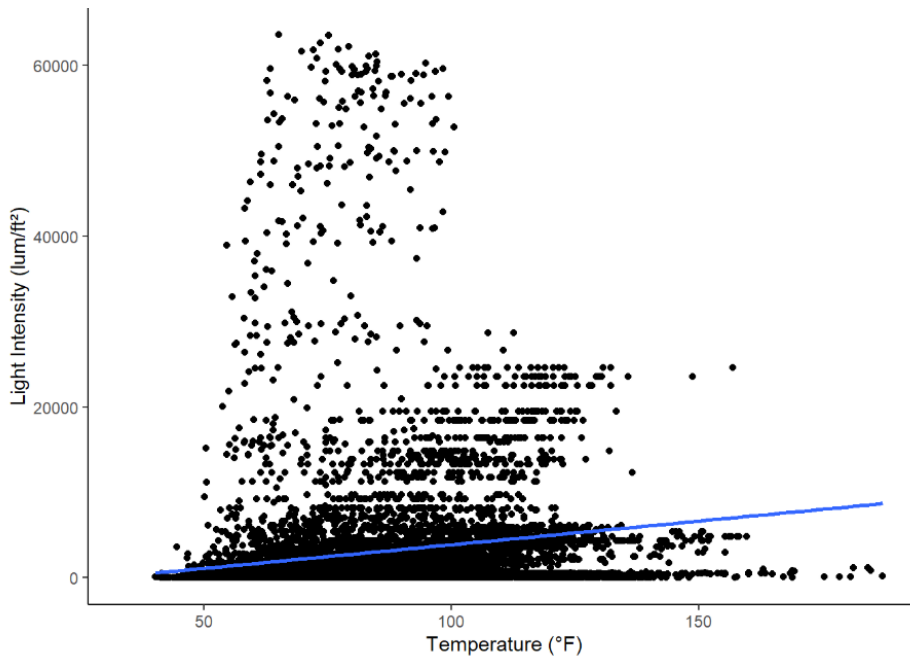
K. Left- General PVC triangular structure and joint. Right-Metal stake and with PVC pipe slid on.



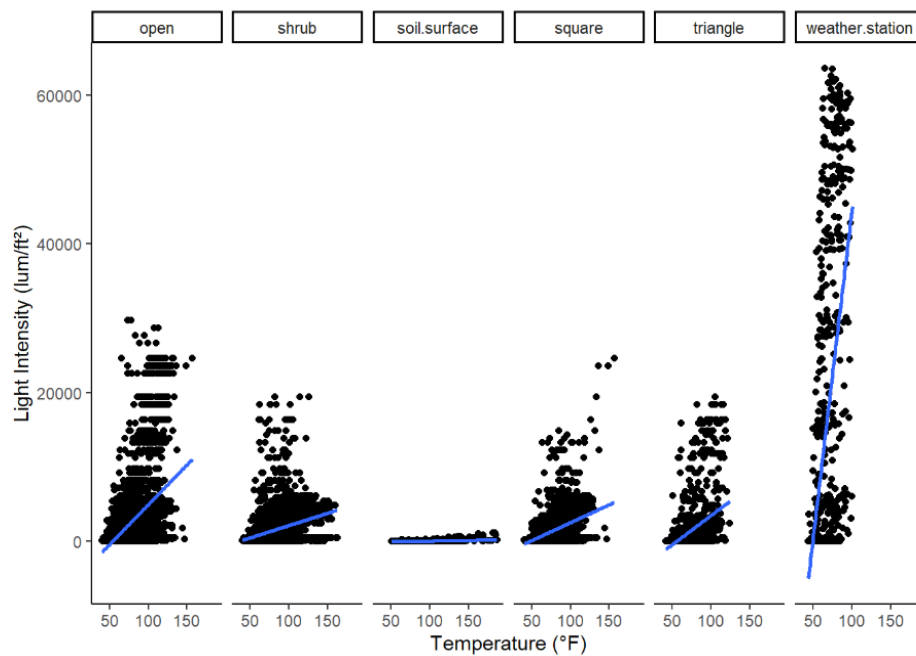
L. Box plot showing ambient temperature (°F) at each microsite/site. Solid middle lines shows the median of the data, whilst whiskers show 1.5 standard deviation. Solid dots are outliers >1.5 interquartile range (IQR). Diamonds dots represent the mean.



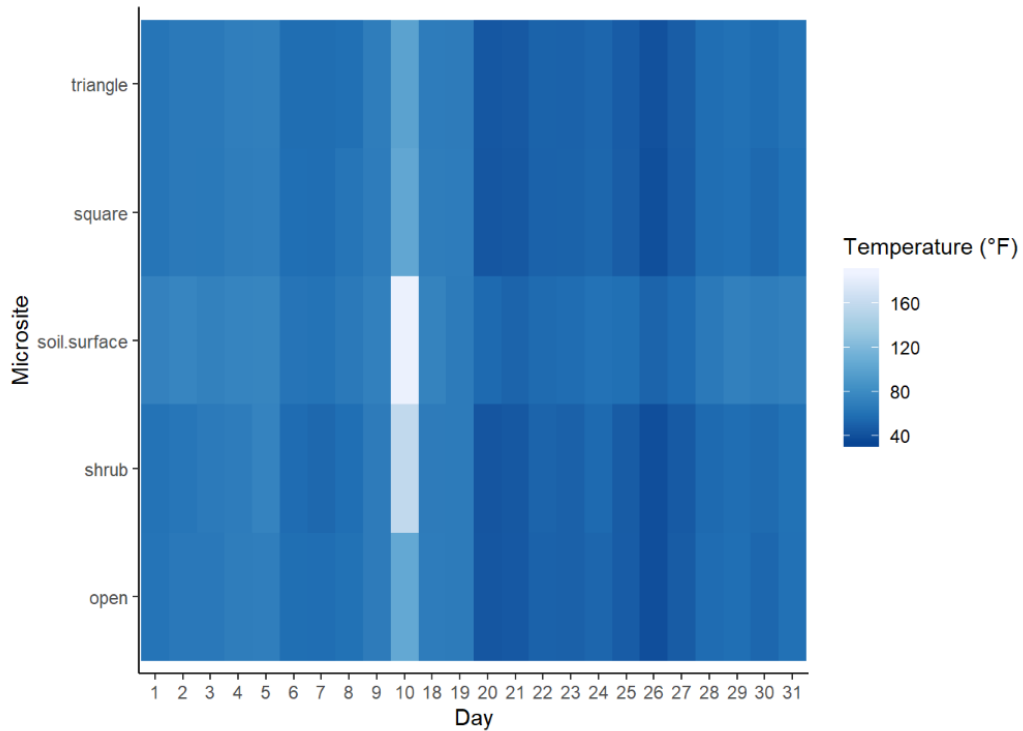
M. Box plot showing light intensity (lum/ft²) at each microsite/site. Solid middle lines shows the median of the data, whilst whiskers show 1.5 standard deviation. Solid dots are outliers >1.5 interquartile range (IQR). Diamonds dots represent the mean.



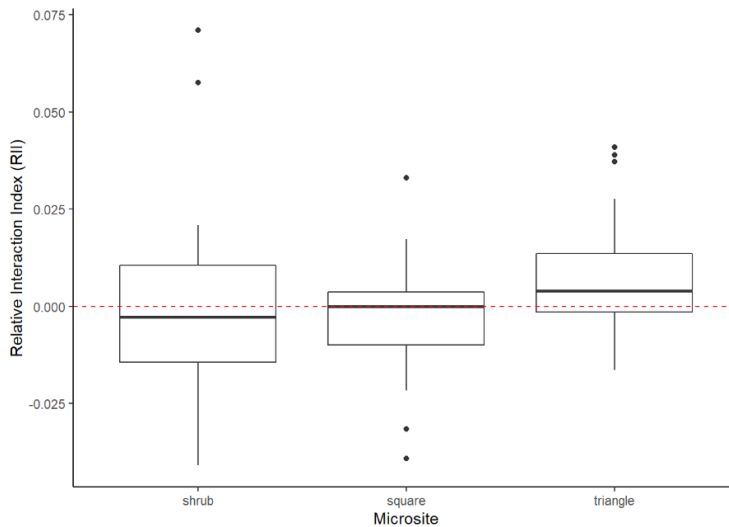
N. Scatterplot showing the overall relationship between solar radiation (lum/ft^2) and temperature ($^{\circ}\text{F}$) (Kendall's $\tau=0.281$, $p=0.0001$). Blue lines represent smooth conditional mean fitted using the method GLM.



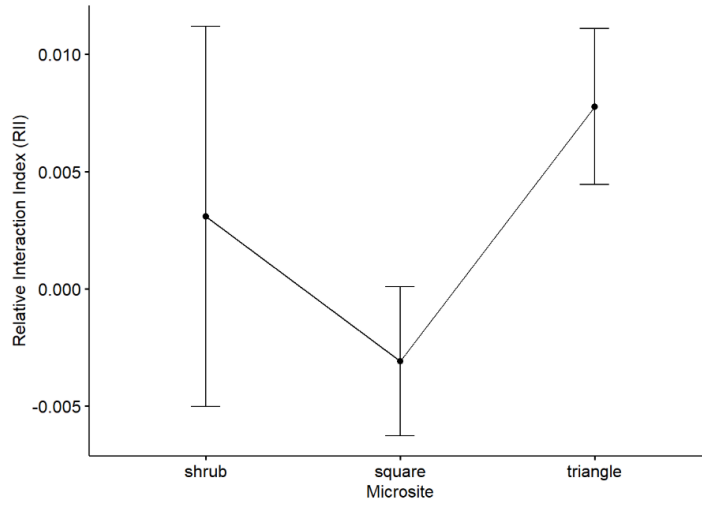
O. Scatterplot showing the relationship between solar radiation (lum/ft^2) and temperature ($^{\circ}\text{F}$) at each microsite. Blue lines represent smooth conditional mean fitted using the method GLM. The relationship is significantly linear at all microsites ($p= 0.0001$), but not for triangle.



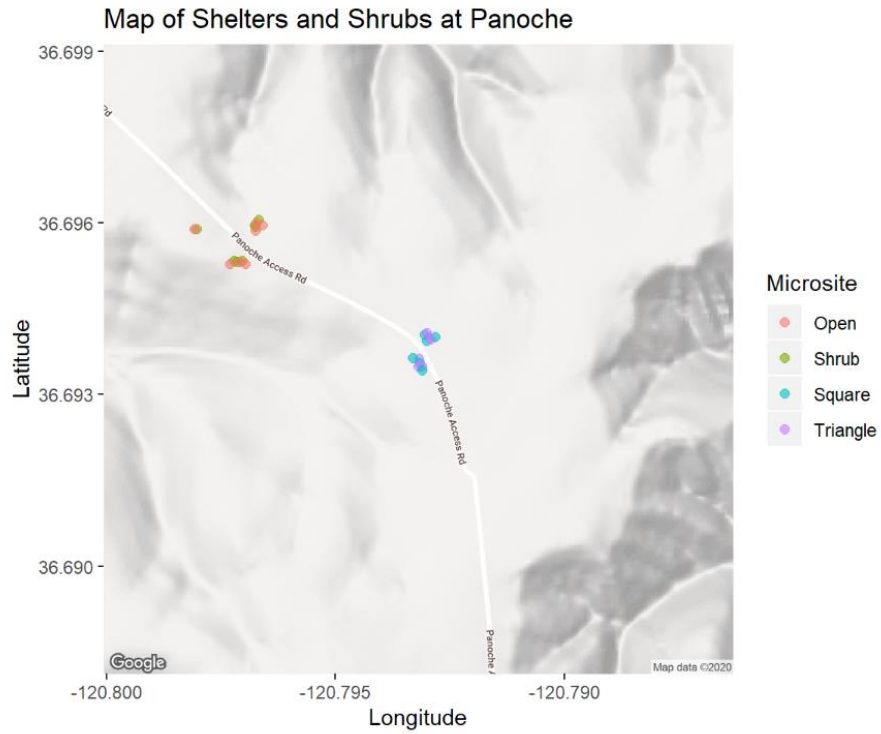
P. Heat Map visualizing temperature (°F) during the study period at the different microsites. Darker blue colours corresponds to cooler temperature whilst bright blue colours correspond to warmer temperatures.



Q. Box plot showing light the Relative Index of Interaction (RII) for each canopy type. Solid middle lines shows the median of the data, whilst whiskers show 1.5 standard deviation. Solid dots are outliers >1.5 interquartile range (IQR). The index values range from -1 to +1. Red dashed line represent 0, or a neutral interaction.

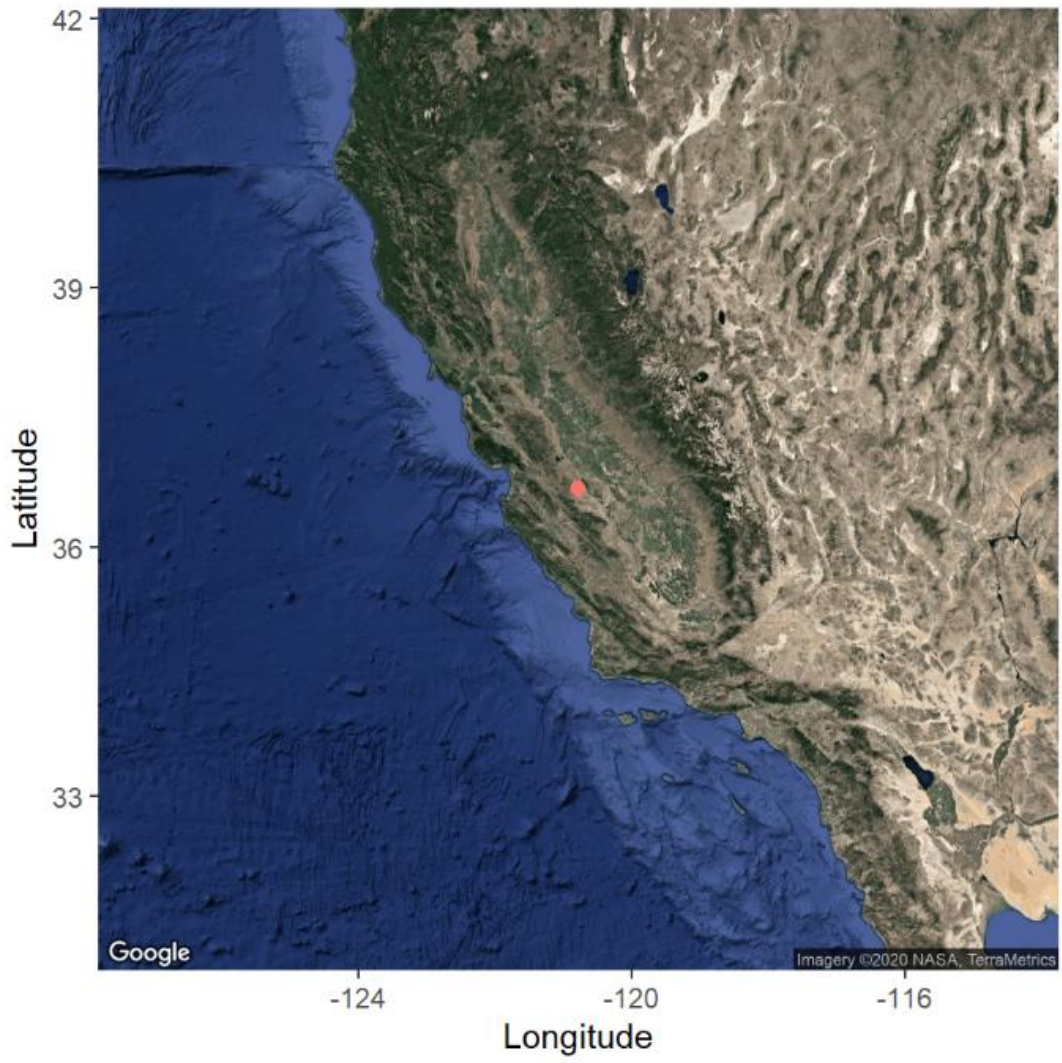


R. Mean point and standard error (SE) plot of RII for each canopies microsite.



S. Map of microsites at Panoche Hills Conservation Area in California.

Map of California



T. Map of site in California, U.S.A.

General Conclusion

Facilitation is the central theme of a large, growing body of literature that explore plant-plant and plant-animal interactions. Abiotic factors have the ability to directly and/or indirectly influence community dynamics. Despite this being a fairly well-known fact amongst ecologists, non-trophic interactions are often forgotten and not incorporated into the dialogue when discussing climate. Furthermore, even when climate is included, most literature often use satellite or weather-station data for their contrast. In this thesis, we used an empirical approach to examine the impacts of artificial shelters on the understory micro-climate. We used micro-climatic data collected via *in situ* loggers and contrasted those against data extracted from the nearby weather station. The mean daily temperatures recorded by the weather-station were significantly lower than microsite level data. Similar results were observed by other studies when contrasting coarser-scale climate against local climate (Lathlean, Ayre, and Minchinton 2011; Kollas et al. 2014). This demonstrates the importance of fine scale climates. Climate can impact behaviour, habitat range shifts, as well as interaction dynamics. Knowing this, it is reasonable to conclude that finer-scale climate is also ecologically important for community-level dynamics. This is not to say that focus should only remain on microsite level data, but instead emphasizes the necessity for micro-macro level contrasts in relevant studies and management practices.

Anthropogenic climate change is perhaps one of the biggest ecological, sociological, and economical threats of the 21st century. Climate change is modifying habitats and living regimes, not just in the western arid regions of the United States, but also worldwide. Species in California face extensive habitat loss as a result of frequent forest fires and rising sea levels. These effects coupled with land-use for urbanization, livestock farming, and petroleum extraction can push many species beyond their range of tolerance and ultimately lead to species loss. Vegetation

canopy plays a key role in how organisms cope with the effects of climate change, yet the re-growth and establishment of natural plants post-disturbance can be extremely slow (Berry et al. 2016), and invasion by non-natives is amongst other confounding factors (Bishop et al. 2019). In this thesis we demonstrated that artificial shelters can function similarly to natural vegetation by lowering the variation in micro-climate and increasing micro-environmental heterogeneity. We showed that by experimenting with different shapes and permeabilities, stake-holders in the region can closely replicate the effects of natural shrubs. In drylands, more species are associated with shrubs than then open because of their facilitative impacts; therefore, a logical next step to this thesis is to test the association of animals with artificial shelters in order to examine whether patterns significantly differ from those of natural shrubs. Camera traps provide us with the means of observing wildlife without human interference (O'Connell, Nichols, and Karanth 2011) that may otherwise influence observations. Using camera trapping methods, we can thus examine association patterns of animals with shelters relative to shrubs, in conjunction with other methods such as field and transect surveys, as well pan trapping and netting. Furthermore, it would be interesting to examine whether there's a significantly-positive relationship between the increase in abiotic stressors, such as light and temperature, and incidents of animals near shrubs and shelters. This can be done using microsite data loggers, such as those used in this thesis, paired with camera traps. These suggested follow-up studies can further deepen our understandings of the utilities of artificial shelters as a form of landscape restoration.

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