View Article Online View Journal

Analyst

Accepted Manuscript

This article can be cited before page numbers have been issued, to do this please use: D. Nodeh-Farahani, J. N. Bentley, L. Crilley, C. B. Caputo and T. C. VandenBoer, *Analyst*, 2021, DOI: 10.1039/D1AN01089A.



This is an Accepted Manuscript, which has been through the Royal Society of Chemistry peer review process and has been accepted for publication.

Accepted Manuscripts are published online shortly after acceptance, before technical editing, formatting and proof reading. Using this free service, authors can make their results available to the community, in citable form, before we publish the edited article. We will replace this Accepted Manuscript with the edited and formatted Advance Article as soon as it is available.

You can find more information about Accepted Manuscripts in the Information for Authors.

Please note that technical editing may introduce minor changes to the text and/or graphics, which may alter content. The journal's standard <u>Terms & Conditions</u> and the <u>Ethical guidelines</u> still apply. In no event shall the Royal Society of Chemistry be held responsible for any errors or omissions in this Accepted Manuscript or any consequences arising from the use of any information it contains.



rsc.li/analyst

8 9 10

11

12 13

14

15

16

₹7

ଞ୍ଚ 8

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59

60

Paper

Received 00th January 20xx, Accepted 00th January 20xx

DOI: 10.1039/x0xx00000x

A boron dipyrromethene (BODIPY) based probe for selective passive sampling of atmospheric nitrous acid (HONO) indoors

Danial Nodeh-Farahani, ^a Jordan N. Bentley, ^a Leigh R. Crilley, ^a Christopher B. Caputo, ^a and Trevor C. VandenBoer* ^a

People spend up to 90 % of their time indoors, and yet our understanding of indoor air quality and the chemical processes driving it are poorly understood, despite levels of key pollutants typically being higher indoors compared to outdoors. Nitrous acid (HONO) is a species that drives these indoor chemical processes, with potentially detrimental health effects. In this work, a BODIPY-based probe was synthesized with the aim of developing the first selective passive sampler for atmospheric HONO. Our probe and its products are easily detected by UV-Vis spectroscopy with molar extinct coefficients of 37863 and 33787 M⁻¹·cm⁻¹, respectively, and a detection limit of 14.8 ng mL⁻¹. When protonated, the probe fluoresces with a quantum yield of 33%, which is turned off upon reaction. The synthesized BODIPY probe was characterized using NMR and UV-Vis spectroscopy. Products were characterized by UV-Vis and ultra high-resolution mass spectrometry. The reaction kinetics of the probe with nitrite was studied using UV-Vis spectroscopy, which had a pseudo-first-order rate of k=7.7 x 10⁻⁴ s⁻¹. The rapid reaction makes this probe suitable for targeted ambient sampling of HONO. This was investigated through a proof-of-concept experiment with gaseous HONO produced by a custom high-purity calibration source delivering the sample to the BODIPY probe in an acidic aqueous solution in clean air and a real indoor air matrix. The probe showed quantitative uptake of HONO in both cases to form the same products observed from reaction with nitrite, with no indication of interferences from ambient NO or NO₂. The chemical and physical characteristics of the probe therefore make it ideal for use in passive samplers for selective sampling of HONO from the atmosphere.

Introduction

Nitrous acid (HONO) is a significant source of hydroxyl (OH) radicals in outdoor environments, the dominant oxidant in the ambient atmosphere. Hydroxyl radicals react with most molecules in the atmosphere, such as volatile organic compounds, to form secondary organic aerosols (SOA) and ozone (O₃). Generally, such reaction products can be detrimental to human and ecosystem health.¹⁻² Although the sources and production mechanisms atmospheric HONO are not completely of understood, current knowledge can be broadly classified into homogeneous gas-phase reactions, heterogeneous surface reactions, and direct emissions.³ Of emerging concern is the role of HONO in impacting indoor chemistry, for which we explore the utility of a BODIPY-based probe to be applied to passive gas sampling in this work.

Similar to outdoors, HONO is suspected to be a major source of OH radicals indoors.^{4–6} HONO can easily photodegrade at low photon energies to form

Electronic Supplementary Information (ESI) available: [details of any supplementary information available should be included here]. See DOI: 10.1039/x0xx00000x

NO and OH via R1 and may dominate oxidation under natural low light conditions indoors.^{7–11}

HONO + hv ($\lambda \leq 400 \text{ nm}$) \rightarrow NO + OH (R1)

Homogenous gas-phase reaction of NO with OH outdoors.8,15 is important for production Heterogeneous reactions form HONO through several pathways.^{12–14} Known HONO production pathways include the hydrolysis of NO₂ on wet surfaces (R2), and light-induced reactions of NO₂ on aerosols and other surfaces. The relative significance of one pathway over another is not yet well constrained.^{15–17} Direct emissions outdoors include vehicle exhaust, biomass burning, and biogenic production from oxidation or reduction of reactive nitrogen by microbes in soils.¹⁸⁻²⁶ Similar loss and production pathways are expected indoors.

 $2NO_2 + H_2O \rightarrow HONO_{(g)} + HNO_{3 (aq)}$ (R2)

To date, HONO levels indoors have been found to be substantially higher than outdoors.²⁷ This is of concern as inhalation of HONO can cause respiratory tract irritation and a decrease in lung capacity and function.^{5,27} Studies to date have measured HONO mixing ratios in indoor

^{a.} Department of Chemistry, York University, Toronto, Ontario M3J 1P3, Canada.

Analyst

Paper

1 2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

₹7

<u>ଅ</u>8

environments with an average near 5 parts per billion by volume (ppbv), while 24-hour average outdoor levels in North America have mixing ratios below 1 ppbv.^{4,6,27–32} In an indoor environment, HONO can be emitted via combustion sources such as candles, gas stoves or cigarette smoke. These sources release HONO in equal amounts to NO_x; reaching tens to hundreds of ppbv mixing ratios which are a concern for inhalation exposure.^{6,28,33} Surface conversion of NO₂ to HONO on hydrated or acidic surfaces is also suspected to be important.^{5,34-} ³⁷ In addition, sources of amines are higher indoors, particularly in smoking households, and the reaction of HONO with secondary and tertiary amines (R3) forms carcinogenic nitrosamines. Thus it is crucial to understand the controls on HONO chemistry and levels in indoor environments.^{38,39} It is also important to develop safe and non-intrusive methods capable of quantifying indoor HONO at ppbv levels or lower.

HONO + (R)₂NH \rightarrow (R)₂NNO + H₂O (R3)

Publicitadour 23 dans 12021. Downloaded by Yack Unikersity 0 6 8 2 9 5 4 8 7 7 1 0 6 8 2 9 5 4 Instruments commonly used for measuring outdoor HONO could be considered unsafe, intrusive, or disruptive to occupants, whose unperturbed activities are necessary for understanding indoor chemical transformations. Such instruments include chemical ionization mass spectrometry (CIMS)^{24,31,40}, long path absorption photometry (LOPAP) ^{14,41,42}, online ion-chromatography systems, and differential optical absorption spectroscopy (DOAS). ^{11,13,43} All require expert operators and have space, power, or safety restrictions that preclude their routine application in indoor environments. Those using toxic chemicals, reagents, and radioactive sources are unsafe or logistically impossible to introduce around occupants in indoor environments to study relevant chemistry driven by 41 their typical activities. Perhaps the most important 42 limits are the costs of purchasing and running one of 43 these instruments, limiting the capability to make 44 simultaneous measurements in many different 45 locations. Obtaining a large statistical pool of 46 observations with research grade instrumentation is 47 therefore challenging, yet crucial, in surveys of 48 indoor air composition as well as exposure studies. 49

Passive sampling provides the bridge between 50 detailed observations and large surveys. They are 51 time-integrated collectors of target gases that 52 diffuse to their surface, followed by reaction or 53 partitioning into the sampling substrate. The passive 54 analyte collection is generally followed by aqueous 55 extraction and offline measurements of the 56 collected molecule(s). Passive samplers are 57 reproducible, silent, cheap, and simple to use.44,45 58 They do not need a power source or any field 59

calibration, making them ideal for use in remote and indoor environments.44,46,47 They selectively sample the analyte of interest through targeted reaction chemistry, yielding a stable analyte in a predictable matrix and a small volume of extraction solvent. Passive samplers combine active air sampling protocols into one step, which simplifies storage and transport. Lastly, they can detect lower quantities of an analyte gas by merely increasing their sampling time.⁴⁵ Due to these advantages, the technique is typical for NO_x ($\equiv NO_2 + NO$) and NO_2 indoor and measurements in outdoor environmental surveys where large numbers of measurements are required.⁴ In this work, we extend this strategy to the sampling of HONO. This comes with the added benefit that HONO measurements allow targeted samplings of nitrogen oxides to be corrected for known positive bias due to HONO interference.⁴⁸ Mitigating this bias is critical to ensuring reliable measurements of these regulated reactive gases indoors. Taken together, these needs justify the development of the first selective passive sampling measurement for HONO.

4,4-Difluoro-4-bora-3a,4a-diaza-s-indacene (BODIPY) are a class of popular fluorescence probes especially in the field of bioimaging for labelling proteins and DNA.^{49,50,51} They are also ideal for use in passive sampling. This is mainly because of their high molar absorptivity coefficients and high fluorescence quantum yields.⁵² However, few water-soluble BODIPY probes have been reported; a property necessary to be of use under environmental conditions or extraction and dissolution into aqueous media.53,54

Since nitrite is present in the human body, wastewater, and many other systems, highly absorptive and/or fluorescent molecular probes for nitrite detection have been reported.55,56 Thus, some BODIPY based probes have been developed to quantify nitrite.⁵⁵ However, attempts to sample and detect its conjugate acid (HONO) has never been done before.^{57,58} This suggests that BODIPY probes could be effective in addressing the current limits of HONO passive sampling in indoor environments, but has not yet been demonstrated. The this development of a selective and sensitive HONO probe that is amenable to aqueous environments and stable for quantitation remains to be explored.

Herein, we synthesized a water-soluble amino-BODIPY probe that reacts with both nitrite and HONO. The optical properties of the molecule were characterized. We quantified the reaction rate of the probe with nitrite and characterized the reaction products for quantification. Proof-ofconcept experiments with calibration gas sources of HONO demonstrate the utility of our probe for application to passive sampling.

3

4

5

6

7

8

9

10

11

12

13

14

15

16

₹7

<u>ଅ</u>8

873/503/203/12:57 873/203/12:57

41

42

43

44 45

46

47

48

49

50

51

52

53

54

55

56

57

58 59 60

Analyst

Experimental

Reagents and instruments

Commercially available chemicals were purchased from either Sigma-Aldrich (Burlington, ON, CA), TCI Chemicals (Portland, OR, USA), or Oakwood Chemicals (Estill, SC, USA) and employed without further purification; unless otherwise stated. In house 18.2 MΩ·cm Milli-Q[®] Water Purification System (Milli-Q[®] Direct 8, Millipore Sigma, MA, USA) were used to obtain deionized water (DIW).

Emission and excitation measurements were recorded with a fluorescence spectrometer (Model FS5, Edinburgh Instruments Ltd, UK). Emission spectra were excited at the respective absorption maxima unless stated otherwise. Ultraviolet-visible (UV-Vis) measurements were recorded with either a Cary 5000 UV-Vis-NIR Spectrometer or Ocean Optics Flame spectrometer FLAME-S-XR1-ES (Gamble Technologies Limited, ON, CA). Recordings were obtained at 25 °C and taken with the instrument operating in single beam mode. Quantum yield measurements were performed by switching the cuvette holder to an SC-30 Integrating Sphere holder (Edinburgh Instruments Ltd, UK) on the Edinburgh FS5 instrument. All measurements were referenced to deionized water, as the solvent. All absorption and fluorescence experiments were conducted in guartz cuvettes (1 cm x 1 cm) sealed with a threaded PTFE cap (Spectrosil®23-Q-10, Starna Scientific, UK).

Experiments monitored by NMR spectroscopy were conducted in NMR spectrum tubes (8" x 5 mm) sealed with standard plastic caps and wrapped with Parafilm or J-young screw cap. ¹H, ¹¹B, ¹³C{¹H}, ¹⁹F{¹H} NMR spectra were acquired at 25 °C on either a Bruker 700 MHz, DRX 600 MHz, ARX 400 MHz, or ARX 300 MHz Spectrometers. Chemical shifts are reported relative to SiMe₄ and referenced to the residual solvent signal (${}^{1}H$, ${}^{13}C{}^{1}H$) of CDCl₃ (δ 7.26, 77.16 μ g/mL) or C₆D₆ (δ 7.16, 128.06 μ g/mL). ¹¹B and ¹⁹F{1H} NMR spectra were referenced relative to 15% BF₃-Et₂O. NMR spectra were analyzed using either TopSpin 4.0.1 (Bruker Corporation, Milton, ON, Canada) or MestReNova 12.0.3-21384 (Mestrelab Research, Spain) software. Data analysis and graphics were generated using WaveMetrics Igor Pro 8.04 (Portland, OR, USA).

Synthesis

A detailed description of the synthesis for the $2,6_{\rm e}$ disodiumsulfonyl-1,3,7,9-tetramethyl $-4^{1039/D1AN01089A}$ aminophenyl)-BODIPY (BODIPY-NH₂), which is a known compound, used in this work is presented in the Supporting information.

NMR for 2,6-disodiumsulfonyl-1,3,7,9-tetramethyl-5-(4-aminophenyl)-BODIPY (BODIPY-NH₂)

The following observations were obtained from the NMR measurements. ¹H NMR (400 MHz, Methanold4) δ 7.00 (d, J = 8.4 Hz, 1H), 6.90 (d, J = 8.4 Hz, 1H), 2.80 (s, 4H), 1.84 (s, 4H). ¹¹B NMR (128 MHz, Methanol-d4) δ 0.73 (t, J = 32.1 Hz). ¹³C NMR (101 MHz, Methanol-d4) δ 155.87, 151.03, 148.42, 143.55, 135.52, 132.35, 130.00, 123.91, 116.60, 14.38, 13.80. ¹⁹F NMR (377 MHz, Methanol-d4) δ -144.46 (dd, J = 64.3, 31.8 Hz). The data was in agreement with literature, confirming that our probe was pure and could be interrogated for its optical and chemical properties.⁵³ Fig 1 shows the synthesis pathway for the BODIPY-NH₂ probe. NMR spectra of BODIPY-NH₂ are shown in Figs S9-12.



Fig 1. BODIPY-NH₂ synthesis scheme

Preparation of BODIPY-NH₂ and nitrite solution for reaction

A stock solution of BODIPY-NH₂ was made by dissolving 3.1 mg of its crystal (NewClassic MS, Mettler Toledo, ON, CA) in DIW using a 100 mL Pyrex volumetric flask. The nitrite (NO_2^{-1}) solution for each experiment was made daily by dissolving 19.6 mg of its sodium salt in DIW using a 50 mL CorningTM Class B polypropylene volumetric flask with tapered polypropylene stopper. This stock solution was diluted to reach a NO_2^{-1} concentration 10 times greater on a mole basis than the BODIPY-NH₂ to carry out reactions. To facilitate the reaction, one drop of concentrated HCl (34-36%) was added to 1.5

3

4

5

6

7

8

9

10

11

12

13

14

15

16

₹7

<u>ଅ</u>8

823/203/12:72 823/203/12:72

41

42

43

44 45

46

47

48

49

50

51

52

53

54

55

56

57

58 59 60 mL of BODIPY-NH₂ solution to generate BODIPY-NH₃⁺.

Sampling gaseous HONO with BODIPY-NH₃⁺

A solution of BODIPY-NH₃⁺ collected HONO at ppbv mixing ratios for three days via bubbling in an unfritted-impinger (CG-1820-01, 04, and -10, Chemglass Life Sciences, Vineland, NJ). The sampled HONO was produced at 100 sccm from a calibration source, according to the method of Zhou et al.⁴ The BODIPY-NH₃⁺was made by diluting 5 mL of 314.8 μ g/mL BODIPY-NH₂ with 5 mL of concentrated HCl and DIW in a 50 mL volumetric flask, then transferring 25 mL of the solution to the impinger. The gas flow exiting the bubbler was diluted with zero air (8301 Series Zero Air Generator, American Ecotech, Warren, RI, USA) to 1000 sccm to meet the flow requirements of a Serinus EC9841 NO_x analyzer (American Ecotech) demonstrated previously to quantitatively measure HONO.4 The bubbler was bypassed to conduct negative controls by supplying zero air and positive controls of HONO from the calibration source. A schematic of the experimental setup is provided in Fig S14.

Mass spectrometry characterization of the reaction product

The reaction product was directly infused into an electrospray ionisation source that was operated in negative ion mode on an Agilent 6538UHD Q-TOF MS (Agilent Technologies, CA, USA). Analysis was performed on three samples: i) 2.5 mL of 314 μ g/mL BODIPY-NH₃⁺as an unreacted sample, ii) 2.5 mL of 314 μ g/mL BODIPY-NH₃⁺ reacted with 0.5 mL of 2360 μ g/mL NaNO₂, and iii) the resulting solution obtained after bubbling HONO in BODIPY-NH₃⁺.

Results and Discussion

Optical properties of BODIPY-NH₂ upon reaction with NO₂-

The absorption maxima of the probe (BODIPY-NH₂), the conjugate acid (BODIPY-NH₃⁺), and the product(s) after reaction with NO₂⁻ were measured by UV-Vis and determined to be 496 nm, 500 nm, and 506 nm, respectively (Fig 2). There is a 10 nm difference between the absorption maxima of BODIPY-NH₂ and the product of the reaction indicating that the product is sufficiently resolved to be detected by UV-Vis in a mixture. The time lar absorptivity coefficients of BODPY 10142, BODPY NH_{3}^{+} , and the product were measured using absorbance profiles from a range of solution concentrations (Fig S15) and were calculated to be 37860 ± 3200 M⁻¹ cm⁻¹ (λ =496 nm, n=4, 1 σ standard deviation), 38200 \pm 3570 M⁻¹ cm⁻¹, (λ =500 nm, n=4, 1σ), and 33790 ± 1900 M^{-1} cm⁻¹ (λ=506 nm, n=4, 1σ), respectively. The BODIPY-NH₂ showed stability over 8 months of storage in aqueous solution which was acidified to produce BODIPY-NH₃⁺ prior to the measurement (Fig S16). When HCl was added to the BODIPY-NH₂ the amino some of the sulfonate groups may have been protonated. Protonation of the nitrogen lone pair converts the donor substituent to an acceptor, which in turn, bathochromically shifts the optical properties and improves the fluorescence quantum yield.

The optical properties of the reaction product are suitable for quantitative analysis where fluorescence detectors are not available or monitoring the loss of BODIPY-NH₂ would be inaccurate. To this end, we forced a reaction of BODIPY-NH₂ to completion with 10 times excess NO_2^- on a molar basis.



Fig 2. Absorbance of 15.5 μ g/mL solutions of BODIPY-NH₂, BODIPY-NH₃⁺, and the product of reaction between BODIPY-NH₃⁺ and excess nitrite measured with Edinburgh FS5.

After reaction with NO_2^- , it was observed that the product colour visible by eye to be different than that of either BODIPY-NH₂ or BODIPY-NH₃⁺. No fluorescence was detected for the reaction product.

3

4 5

6

7

8

9

10

11

12

13 14

15

16

₹7

<u>ଅ</u>8

\$7350312:57: 2010 00 27:37:1

23

Wownloaded by Mark Winkersty

Publishedron 23 August 2021.

41

42

43

60

Quantum yields for photons emitted between 475-700 nm for the BODIPY-NH₂, BODIPY-NH₃⁺, and product of the reaction, are 1% (λ_{ex} =495 nm), 33% (λ_{ex} =495 nm), and 0% (λ_{ex} =512 nm), respectively (see Fig S17 for emission spectra). Since the BODIPY-NH₃⁺ can be clearly discerned by its intense emissive properties, it shows that this molecule could act as a fluorescent ON-OFF sensor for NO₂⁻ detection by measuring the fluorescence decay signal. The molar absorption coefficient of the product remaining very high means the use of BODIPY-NH₂ as a sensor could detect low quantities of NO_2^- and/or HONO. The best quantitative performance would be expected where the product can be separated from the probe, prior to detection by UV-Vis. To gain insight into how this might be carried out, we analysed a sample of the reaction product to determine its sensitivity in our simple UV-Vis system and then by mass spectrometry to discern its structure.

Sensitivity of the probe

The sensitivity of the probe was determined by analysing a series of varying concentrations of the reaction product. The absorbance was measured with our UV-Vis Ocean Optics Flame spectrometer with a 1 cm path length. Measured absorbance was plotted against concentration (Fig S18) and the limit of detection (LOD) calculated as three times the standard error of the slope divided by the slope. Based on a 1:1 mole ratio between reaction of BODIPY-NH₃⁺ and HONO, the LOD for the probe was determined to be 14.8 ng mL⁻¹ of HONO in solution. We note that this calculated LOD is specific to this instrument and would improve by increasing the path length and spectral resolution.

Characterization of the products of the reaction of probe with HONO and NaNO₂

44 The products of the reaction of BODIPY-NH₃⁺ with 45 excess NaNO₂ in acidic solution were detected at 46 483.07, 499.06, and 517.03 amu. For the reaction of 47 BODIPY-NH₃⁺ with the gas-phase HONO (see below), 48 the same three products were observed, suggesting 49 a similar pathway for both reactions. The detection 50 of products with m/z 499.06 and 517.03 is 51 consistent with the reactions proceeding via 52 nitrosonium formation. The protonation of NO_2^- to 53 HONO is followed by the reaction of nitrosonium 54 and BODIPY-NH₂ to form the nitrosamine, and 55 finally, this forms a diazonium salt. The diazonium 56 salt is unstable and readily loses N_2 and the 57 molecule undergoes additions as shown in Fig 3. 58 This chemistry is well established and exploited in 59

other reactions that detect NO₂⁻ in environmental samples.^{56,59} DOI: 10.1039/D1AN01089A



Fig 3. Proposed pathway from the reaction of BODIPY probe with nitrite and HONO to products with m/z of 499.06, and 517.03 amu

These products were confirmed with mass spectrometry, and the presence of the boron atom in our BODIPY probe allows the use of isotopic patterns when isolating the reaction product from the sample solution. Three samples: a protonated BODIPY-NH₃⁺ as an unreacted sample, the BODIPY-NH₃⁺ reacted with excess NaNO₂, and the BODIPY- NH_3^+ bubbled with gaseous HONO were analysed. The collected mass spectra are provided in Figs S19-21. The three ions with mass to charge ratios (m/z) of 499.06, 517.03, and 483.07 contained the boron isotopic pattern indicating the presence of part of the BODIPY core in the reaction products. The ion with m/z 499.06 corresponds to the presence of a hydroxyl group in place of the amino on the BODIPY- NH_2 probe. The ion with m/z 517.03 corresponds to the presence of a chlorine substituent, and at m/z 483.07 a hydrogen in place of the amino on the BODIPY-NH₂, respectively.

The unreacted sample had two ions present in the mass spectra: a dominant peak at m/z 498.08 corresponding to BODIPY-NH₂ (Fig S19), and another peak at m/z 483.07 with roughly 40 times lower intensity. Sodium and chloride adducts of BODIPY-NH₂ were also observed.

The products of the reaction of BODIPY-NH₃⁺ with excess sodium nitrite had three abundant ion peaks in the mass spectra at m/z of 483.07, 499.06, and 517.03 with relative intensities of 26%, 27%, and 47%, respectively. The product of BODIPY-NH₃⁺ bubbled with gaseous HONO had the same three

St Accepted Manu

Paper

1

peaks in its mass spectra at m/z of 483.07, 499.06, and 517.03 with relative intensities of 43%, 22%, and 35%, respectively. In both product samples, there was no peak detected corresponding to the BODIPY-NH₂ probe at m/z of 498.08, meaning that the reaction in both samples went to completion, or at least far enough so as to be undetectable by this instrumental technique. In order to provide extra assurance that the ions detected correspond to the structures shown in Fig 4., the ions at m/z 483.07, 499.06, and 517.03 were selected for high-resolution accurate mass analysis and processed with the molecular formula software algorithm in MassHunter Qual B.06.00 (Agilent Technologies, CA, USA) to confirm the atomic formulas.



Fig 4. The products of the BODIPY-NH₂ reaction with nitrite and HONO determined from high resolution accurate mass analysis with m/z of 483.07, 499.06, and 517.03 amu from left to right

Kinetics of the reaction

We investigated the rate of the reaction between our BODIPY-NH₂ and NO_{2⁻} in acidic aqueous solution. This was carried out under pseudo-first-order conditions with 10 times molar excess of nitrite. Spectra were acquired every minute over the 10minute reaction time. The first spectrum was gathered immediately after the NO₂⁻ was added to the BODIPY-NH₃⁺ solution. A shift in maximum absorption was observed from 500 nm to 506 nm. This indicates that the product(s) of reaction between NO_2^- and BODIPY-NH₃⁺ formed and the reaction had proceeded to completion (Fig 5). Which of the three, or combination thereof, detected by MS dominates in these experiments was not possible to discern and will be the focus of future work. The reaction was complete in under 10 minutes. An isosbestic point at 505 nm shows the transition from BODIPY-NH₃⁺ to product(s) happened without forming any long-lived intermediates. Assuming a pseudo-first-order reaction, the rate of reaction could be calculated (k=7.7x10⁻⁴ s⁻¹; Fig S22), resulting in an upper limit for the second order rate of 1.33 M⁻¹ s⁻¹ for our experimental concentration of nitrite. The rate of reaction is fast enough that even at a 1:1 mole ratio of NO_2^- to the probe, the reaction would to be complete in minutes. To vapply any sampling probe to the collection of HONO from indoor air, such a fast and effective reaction is required. This is particularly true for passive sampling where collisions between the analyte gas and the probe molecule fixed to the surface initiate the reaction sequence. The rapid rate of reaction observed in solution provided sound evidence that our probe was suitable to test for its ability to react with gas-phase HONO from a calibration source.



Fig 5. Changes in absorbance of 15.5 mg L⁻¹ BODIPY-NH₃⁺ in acidified aqueous solution with 10 times molar excess nitrite at pH 0, recorded at 1-minute intervals over 10 minutes measured with Edinburgh FS5.

Sampling gaseous HONO with BODIPY-NH₃⁺

Initial testing of BODIPY-NH₃⁺ to sample gaseous HONO was carried out using atmospherically relevant mixing ratios (2-4 ppbv). The HONO was produced from a calibration source as described in Lao et al.⁶⁰ The HONO was sampled for eleven days so the reaction could reach completion. When HONO was bubbled into the BODIPY-NH₃⁺ solution, the mixing ratio decreased dramatically (90-100%). This suggested that HONO reacted rapidly with the available BODIPY-NH₂ in the acidified solution.

HONO is expected to form the nitrosonium cation in acidic solutions, as shown in Fig 3. Thus, 1 M HCl was selected as a non-oxidizing source of hydronium ion, enabling efficient conversion of HONO to nitrosonium at a pH near zero. The nitrosonium cation reaction with aromatic amines would follow the diazonium pathway and form some of the products observed by MS (Fig 3).⁵⁹ An advantage of

Analyst

Paper

Analyst

sampling into aqueous solution with a pH near zero is to maintain the protonation of the amino group on the BODIPY-NH₂. This limits the reactivity of the amino group toward addition at electrophilic sites and minimizes the decomposition of the BODIPY probe.

The reaction of BODIPY-NH₃⁺ with HONO (Fig 3) was confirmed by the observation of a shoulder in the UV-Vis product spectrum (Fig 6). Unexpectedly, the decreased absorbance observed by UV-Vis indicated either decomposition or physical loss mechanisms were present. We expected the absorption maxima to shift (as observed in Fig 5), but not that the integrated absorbance would decrease (Fig 6). This is because the molar extinction coefficients of the precursor and product were determined to be similar (~35000 M⁻¹cm⁻¹). We hypothesized that the additional loss could be due to the apparatus (e.g. the sintered glass frit of the impinger) and/or physical removal of the probe in the form of aerosols which could be generated by bubbles bursting. To explore this loss mechanism further, we performed a negative control experiment in which we bubbled a solution of the probe with only ultrapure nitrogen gas at the same flow for eight days. We observed a 30% loss in absorption (Fig S23), suggesting that during this first HONO sampling experiment (Fig 6) the probe was lost by mechanism(s) attributable some to the experimental setup alone. This prevents reaction characterization by a mass balance approach (see below). Therefore, to minimize loss of the probe in future experiments we used a non-fritted impinger and reduced the sampling time by utilising higher HONO mixing ratios. Modification of the HONO calibration source to generate these higher mixing ratios was achieved with 1 g of dispersed sodium nitrite salt instead of a coated sodium nitrite reactor, according to Zhou et al.⁴

To differentiate between the different uptake mechanisms possible in our sampling setup, a series of control experiments were designed in which HONO was bubbled through i) deionized water (DIW), ii) 1 M HCl, and iii) 4.96 μ g mL⁻¹ aniline solution in 1 M HCl (Fig 7). Aniline was chosen because of the similarity of the reaction of aromatic amines with HONO to form diazonium salts. The NO, NO₂, and HONO temporal trends in these experiments were recorded at 1-minute intervals using a modified NO_x analyzer (see methods).



Fig 6. Changes in absorbance of the BODIPY-NH₃⁺ before and after bubbling with 3.5 ppbv of HONO for a period of 11 days measured with Edinburgh FS5.



Fig 7. Time series of NO_x measurements from the exit of the impinger when gaseous HONO is bubbled into deionized water (DIW, pink), 1 M HCl (green), and 4.96 µg mL⁻¹ aniline in 1 M HCl (blue). (A) The fraction of NO generated relative to HONO sampled from the calibration source and (B) the fraction of measured HONO + NO_2 relative to the same HONO input. Curve inflections indicate equilibration,

Paper

1 2

3

4

5

6

7

8

9

10

11

12

13

14

15

16

₹7

8 8/3/2047 2:72 0 8 8/3/2047 2:72 0 8 8/3/2047 2:72 0 8 8/3 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 0 8 9/2047 2:72 0 8 9/2047 0 9/20047 0 9/20047 0 9/20047 0 9/20047 0 9/20047 0 9/2000 9/2000000000000000000

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59 60 breakthrough, and/or production of the detected gases.

From Fig 7B, similar temporal trends in HONO + NO₂ output relative to the HONO delivered were observed when the impinger contained 1 M HCl and DIW, suggesting that there was little difference in the chemical uptake of HONO between these solutions (R4). As expected, there is less holding capacity for HONO in these solutions in the absence of chemical reaction with either aniline or BODIPY-NH₃⁺ and the HONO would start to decompose in solution into NO and NO₂ through R5.

$$HONO_{(aq)} + H_2O \leftrightarrow H_3O^+_{(aq)} + NO_2^-_{(aq)}$$
(R4)

$$2\text{HONO}_{(aq)} \rightarrow \text{NO}_{(g)} + \text{NO}_{2(g)} + \text{H}_2\text{O}$$
(R5)

Aniline was expected to undergo the same diazotization pathway as proposed for our BODIPY- NH_3^+ probe when bubbled with gaseous HONO according to R6:

 $C_6H_5NH_2 + HONO_{(aq)} \rightarrow C_6H_5OH + H_2O + N_{2(g)}$ (R6)

A different trend in the ratio of the HONO + NO_2 exiting the impinger solution compared to HONO input was observed for the aniline solution (Fig 7B). The fraction was persistently low (<10%) for the first 2000 mins (Fig 7B). The absence of HONO and NO₂ in the outflow of the bubbler containing the acidic aniline solution indicates quantitative and irreversible chemical loss, as expected for this positive control. After approximately 2200 minutes of sampling, the fractions of both HONO + NO₂ and NO in the outflow of the bubbler were observed to increase (Fig 7), pointing to a reduction in available aniline for reaction with HONO. At this point, 1.27(±0.38)x10⁻⁶ moles of HONO had been sampled, which was 95±30% of the available aniline (1.33x10⁻⁶ moles) in the solution, consistent with the error resulting from the 10-30% precision of the modified HONO source.⁶⁰ This indicates that the reaction of HONO and aniline (R6) was near or beyond completion and that the experimental setup was valid for mass balance experiments. A detailed treatment of the mass balance calculation is available in the Supporting Information (Section S2). In all control experiments, a steady state in the levels of NO and NO₂ were observed in the air exiting the bubbler solution. In the negative controls, this was reached once they had equilibrated, or in the positive control, when the aniline was reacted to completion (Fig 7). The earlier observation of this rise in NO and NO₂ levels was due to the reduced chemical uptake for HONO (R6) in 1 M HCl and DIW. Thus, the sum of measured NO and

HONO + NO_2 (in the absence of R6) should be equal HONO delivered tool: the 39/Solution. to the Unexpectedly, the sum of NO and NO₂ only reached 90 % of the HONO input in all three control experiments, indicating an additional loss mechanism for HONO that did not generate NO or NO₂. Since the observation was reproducible and only observed in the absence of a reactive sink in our positive control, it was possible to proceed with mass balance reaction experiments with the BODIPY-NH₂.

An aqueous solution of BODIPY-NH₂ containing 1 M HCl sampled gaseous HONO (25 ± 14 ppbv) from our calibration source over three days (to minimize loss of the probe, see above). Fig 8 shows the absorbance change during sampling, while a time series of the HONO uptake as measured by the NO_x analyzer is given in Fig 9.

The observed UV-Vis absorption spectra for this experiment (Fig 8) had a similar pattern to our previous experiment with lower HONO mixing ratios (Fig 6, 2-4 ppbv). The same shoulder was observed at higher wavelengths with increasing HONO sampling duration. A notable difference was that the total absorbance in Fig 8 was not markedly reduced after sampling HONO for 61 hours.



Fig 8. Change in the absorbance of $BODIPY-NH_2$ in 1 M HCl over time when bubbled with 25 ± 14 ppbv of HONO measured with Ocean Optics Flame spectrometer.

As the sample probe loss was minimal, we then proceeded to calculate a mass balance of sampled HONO relative to the amount of probe present in solution, following the same calculation that was validated with our aniline positive control. Briefly, the HONO input to the bubbler was assumed to be the reading the NO₂ channel of during measurements of the calibration source output (E1). Our custom HONO calibration source has previously been shown to only have NO impurities, which are

3

4

5

6

7

8

9

10

11

12

13

14

15

16

₹7

8 8/3/2047 2:72 0 8 8/3/2047 2:72 0 8 8/3/2047 2:72 0 8 8/3 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 2:72 0 8 9/2047 0 8 9/2047 2:72 0 8 9/2047 0 9/20047 0 9/20047 0 9/20047 0 9/20047 0 9/20047 0 9/2000 9/2000000000000000000

Published ou 23 A usus (2021.

41

42

43

44

45

60

Analyst

with

within 10% of the output, and that HONO is quantitatively detected by this instrumental setup.⁶⁰ The uptake of HONO was then calculated (E2). A detailed treatment of the mass balance calculation is available in the Supporting Information (Section S2).

 $HONO_{in} = HONO_{cs}$ (E1)

$$HONO_{uptake} = HONO_{in} - [$$

$$(NO_{output} + (HONO + NO_{2 output})) - NO_{cs}]$$
(E2)

Where HONO in refers to HONO input to the BODIPY-NH₂ sampling solution and HONO _{cs} is the HONO measurement of the calibration source on the NO_x analyzer. The NO _{output} and HONO + NO₂ _{output} are NO and HONO + NO₂ measurement exiting the impinger, as measured by the NO_x analyzer, respectively. HONO _{uptake} refers to the HONO that is lost to the sampling solution in the reaction of probe with HONO.

The uncertainties of the calculated uptake quantity are dominated by the precision of the HONO output from the calibration source when fitted with a salt bed, which can vary in precision from 10-30% in previous work.60 We used the maximum of this range as an upper limit to estimate the uncertainty in the HONO calibration source output. Any missing mass in the nitrogen mass balance was assumed to correspond to the formation of N₂ via diazotization⁵⁹ and the previously discussed decomposition pathway. Identifying the reaction endpoint is challenging, as the exact point where NO and NO₂ produced from R5 reach steady state can be masked by equilibrium processes, which increased the uncertainty. To reduce this, we selected the time where the NO₂ measurement in the output began to increase as an indication of reaction being near completion.

46 From Fig 9, the uptake of gas-phase HONO by the 47 probe was quantitative initially (0-800 mins) as 48 indicated by measured HONO + NO_2 exiting the 49 bubbler being less than 0.04 of the HONO input. At 50 approximately 800 min (Fig 9), the NO₂ and NO 51 exiting the bubbler were observed to rise. This 52 suggests that the reaction of the probe with HONO 53 was near completion and any unreacted was being 54 observed exiting in the bubbler gas flow. At 900 min, 55 through integration of E2 over time, we calculated a 56 total of 1.36(±0.40)x10⁻⁶ moles of HONO sampled, 57 corresponding to 95±30% of the total amount of 58 BODIPY probe in the solution $(1.43 \times 10^{-6} \text{ moles})$, 59

again within the expected 10-30% precision of the modified HONO source.⁶⁰ This is in agreement with the mass spectrometry results (Fig S20) in which no ions at m/z 498.08 corresponding to BODIPY-NH₂ mass were detected. This therefore suggests that the probe was fully converted to products at the end of the experiment.

Once the reaction of our probe was complete, the HONO bubbled into solution either passes through or follows R4 or R5. Due to the reaction matrix consisting of 1 M HCl (pH=0), R5 is expected to be a negligible pathway and HONO loss should instead be dominated by R6. After 2000 min (Fig 9), the impinger output reached a steady state with the levels of NO and HONO + NO₂ at 0.2-0.3 and 0.4-0.5 fraction of HONO input, respectively, and consistent with the observations of our control experiments.



Fig 9 Time series of measured HONO+ NO_2 and NO in the impinger output as a fraction of the HONO input when 31.5 µg mL⁻¹ BODIPY-NH₂ in 1 M HCl solution sampled gas-phase HONO (25±14 ppbv). NO measurement was corrected for the NO impurities from calibration source. Gaps in the time series are due to background measurements. Dashed line at 900 min indicates the time used for mass balance calculation.

The above experiments were performed under controlled conditions (i.e., using zero air) and therefore we next examined the ability of the probe to sample HONO in a matrix of indoor air. For this experiment, the indoor air was first scrubbed of any ambient HONO using a denuder coated with Na₂CO₃ and then HONO of a known mixing ratio (17± 0.2 ppbv) was added to the indoor air using the calibration source prior to sampling with the probe (31.5 μ g mL⁻¹ solution of BODIPY-NH₂ in 1 M HCl). This was to ensure a consistent known amount of HONO was added during this experiment and so that mass balance calculations could be utilized, while not excluding potential ambient interferences such as NO₂ (present at 10-20 ppbv during the experiments, R2). The sampling setup is shown in Fig S24. Fig 10 demonstrates that in indoor air HONO

Analyst Accepted Manuscrip



was taken up by the solution, with the observed change in absorbance similar to the control experiments (Fig 8).



Fig 10. Change in the absorbance of $BODIPY-NH_2$ in 1 M HCl over time when bubbled with 17 ± 0.2 ppbv of HONO added to acid-free room air measured with Ocean Optics Flame spectrometer.

Fig 11 presents the time series of measured NO_x from the output of the BODIPY-NH₂ in 1 M HCl solution during the indoor air sampling. For this experiment the mass balance calculation has higher uncertainty compared to the controlled experiments due to the presence and variability of NO_x levels in the sampled indoor air. However, in the absence of a combustion source, most of the NO_x indoors will be NO₂ and therefore the increase in NO in the outflow of the sampling solution was indicative that the reaction was near completion. This change in NO was observed after approximately 1200 minutes (20 hours) of sampling (Fig 11). The reaction nearing the stoichiometric endpoint is further confirmed by the measured change in the UV-Vis absorbance at this time (Fig 10, 20 hours). Overall, Figs 10 and 11 demonstrate that a solution of BODIPY-NH₂ in 1 M HCl can sample HONO in the presence of matrix components typically found in a sample of ambient indoor air. This specifically suggests that cross-reactivity in the presence of 10-20 ppbv NO_x , mostly as NO_2 , is undetectable in these experiments and therefore minor in nature.



Fig 11 Time series of measured HONO+ NO₂ and NO in the impinger output as a fraction of the HONO input when 31.5 μ g/mL BODIPY-NH₂ in 1 M HCI solution sampled gas-phase HONO (25±14 ppbv). NO measurement was corrected for the NO impurities from calibration source. Dashed line at 1200 min indicates the time used for the absorbance measurement.

Conclusions

In this work, we present a sensitive boron dipyrromethene based probe (2,6-disodiumsulfonyl-1,3,7,9-tetramethyl-5-(4-

aminophenyl)-BODIPY; BODIPY-NH₂) for sampling gaseous HONO. The synthesis of the BODIPY-NH₂ probe is described and confirmed with NMR spectroscopy. The BODIPY probe reactivity was first tested using an acidic solution and sodium nitrite in excess as a source of in-situ nitrous acid. The reaction proceeded readily to completion on a timescale of 10 minutes. Characterization of the products from this reaction by high resolution mass spectrometry indicated diazotization to be the likely reaction mechanism. The probe and the products all displayed a high molar extinction coefficient, with the probe being fluorescent in acidic aqueous solution (Φ =33%) which turns off after reaction with HONO (Φ =0%; see Fig S17). These optical properties indicate viable use of UV-Vis spectrometry to measure HONO via this probe or its products at atmospherically relevant concentrations in the indoor environment (i.e. ppbv range) with high sensitivity.

pseudo-first-order The demonstrated fast reaction kinetics (k=7.7x10⁻⁴ s⁻¹) between the BODIPY probe and HONO, without any measurable intermediates, indicates its suitability for passive sampling in the ambient atmosphere. The rate is not so fast as to create a diffusion limitation in gas sampling, which is ideal for avoiding established issues in target gas collection rates (e.g. air movement from wind or HVAC). We demonstrated that the BODIPY probe would also react quantitatively with gaseous HONO. Mass balance

5

6

7

8

9

Published ou 23 A usus (2021.

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59

60

1

calculations between the amount of probe available and the amount HONO sampled further indicated quantitative uptake and was confirmed by a series of control experiments.

Overall, this work has provided proof-of-concept for using the BODIPY-NH₃⁺ probe to quantitatively measure gaseous HONO concentrations in ambient air. The use of this probe overcomes current 10 limitations in the indoor air quality and exposure 11 passive sampling community. First, we expect the 12 high molar absorptivity enables sensitive detection 13 of HONO at ambient levels in reasonable 14 deployment periods (e.g. days). Second, it is 15 amenable to quantitative detection by many 16 existing analytical instrumentation strategies. Since ₹7 the suite of product molecules have different <u>ଅ</u>8 hydrophobicity, the potential of this probe could be 873503702772 873570377227 maximized by combining it with separation techniques. Last, it is selective for the collection of HONO, allowing it to be quantified, and future NO_x measurements by passive techniques to be corrected for the known interference presented by Downloaded by Yark Wrikersity 7 1 0 6 8 2 9 9 5 7 HONO. The probe could therefore be coated on reactive pads and used in diffusive passive samplers for indoor air quality studies. This will be the focus of future work.

Conflicts of interest

There are no conflicts to declare.

Author Contributions

TV and CC supervised the experiments, acquired the funding, and provided resources to support this work. TV JB and CC conceptualized the molecular probe. JB synthesized and characterized the probe by NMR. DNF conducted all reactions and characterizations by UV-Vis, UHRMS, and with the NOx analyzer, including data analysis and visualization. LC aided in the setup of gas sampling experiments and data analysis. DNF created the schematics and reaction mechanisms and prepared the manuscript with contributions from all authors. All authors participated in data analysis and the review and editing of the manuscript.

Acknowledgments

Financial support for this work was provided by the Alfred P. Sloan Foundation Chemistry of Indoor Environments Program (G-2018-1051). TCV and CBC declare research support by NSERC Discovery

Grants (RGPIN-2020-06166 and RGPIN-2018-05827 and TCV by the Early Career Researcher9Discovery Launch Supplement (DGECR-2020-00186). We thank Howard N. Hunter for assistance with the NMR measurements. We thank Dr. Thomas Baumgartner for access to and help with the spectroscopic measurements Edinburgh instruments. We acknowledge the help of Dr. Joshua Gaffen and Amy Laturski with setting up and operating the Edinburgh instruments. We thank Carolyn E. Hempstead for the loan of the Ocean Optics spectrometer and Melodie Lao for their help with the HONO calibration source.

Notes and references

4

- 1 Y. B. Lim and P. J. Ziemann, Environ. Sci. Technol., 2009, 43, 2328-2334.
- A. T. Lambe, T. B. Onasch, D. R. Croasdale, J. P. Wright, A. T. Martin, J. P. Franklin, P. Massoli, J. H. Kroll, M. R. Canagaratna, W. H. Brune, D. R. Worsnop and P. Davidovits, Environ. Sci. Technol., 2012, 46, 5430-5437.
- З L. R. Crilley, L. J. Kramer, B. Ouyang, J. Duan, W. Zhang, S. Tong, M. Ge, K. Tang, M. Qin, P. Xie, M. D. Shaw, A. C. Lewis, A. Mehra, T. J. Bannan, S. D. Worrall, M. Priestley, A. Bacak, H. Coe, J. Allan, C. J. Percival, O. A. M. Popoola, R. L. Jones and W. J. Bloss, Atmos. Meas. Tech., 2019, 12, 6449-6463.
 - S. Zhou, C. J. Young, T. C. VandenBoer, S. F. Kowal and T. F. Kahan, Environ. Sci. Technol., 2018, 52, 8355-8364.
 - S. Gligorovski, J. Photochem. Photobiol. A Chem., 2016, **314**, 1–5.
- 6 V. Bartolomei, E. Gomez Alvarez, J. Wittmer, S. Tlili, R. Strekowski, B. Temime-Roussel, E. Quivet, H. Wortham, C. Zetzsch, J. Kleffmann and S. Gligorovski, Environ. Sci. Technol., 2015, 49, 6599-6607.
- C. J. Young, S. Zhou, J. A. Siegel and T. F. Kahan, Environ. 7 Sci. Process. Impacts, 2019, 21, 1229–1239.
- J. Kleffmann, ChemPhysChem, 2007, 8, 1137-1144. 8
- R. Volkamer, P. Sheehy, L. T. Molina and M. J. Molina, 9 Atmos. Chem. Phys., 2010, 10, 6969-6991.
- 10 S. Kim, T. C. VandenBoer, C. J. Young, T. P. Riedel, J. A. Thornton, B. Swarthout, B. Sive, B. Lerner, J. B. Gilman, C. Warneke, J. M. Roberts, A. Guenther, N. L. Wagner, W. P. Dubé, E. Williams and S. S. Brown, J. Geophys. Res. Atmos., 2014, 119, 6886-6896.
- 11 C. J. Young, R. A. Washenfelder, J. M. Roberts, L. H. Mielke, H. D. Osthoff, C. Tsai, O. Pikelnaya, J. Stutz, P. R. Veres, A. K. Cochran, T. C. Vandenboer, J. Flynn, N. Grossberg, C. L. Haman, B. Lefer, H. Stark, M. Graus, J. De Gouw, J. B. Gilman, W. C. Kuster and S. S. Brown, Environ. Sci. Technol., 2012, 46, 10965-10973.
- J. Zhang, J. Chen, C. Xue, H. Chen, Q. Zhang, X. Liu, Y. Mu, Y. 12 Guo, D. Wang, Y. Chen, J. Li, Y. Qu and J. An, Sci. Total Environ., 2019, 681, 110-123.
- J. Stutz, H. J. Oh, S. I. Whitlow, C. Anderson, J. E. Dibb, J. H. 13 Flynn, B. Rappenglück and B. Lefer, Atmos. Environ., 2010,

Analyst

Analyst Accepted Manuscript

Λ	n	2	ь	10	÷
\sim		а	11	73	1

ו כ	
2	
4	
5	
6	
7	
8	
9 10	
11	
12	
13	
14	
15	
8	
9	
<u></u> 20	
21 221	
∦∠ ∋33	
₹3 24	
25	
<u>-</u> 26	
Ž7	
28	
79 29	
30 1 1 1 1	
a 32	
<u>র</u> ं3	
3 <u>3</u> 4	
3 5	
36	
ವಿ/ ವಿ/	
39	
2 2 40	
41	
42	
43	
44 45	
46	
47	
48	
49	
50	
51 52	
52	
54	
55	
56	
57	
58 50	
60	

Paper

44 , 4090–4098.	33	Government of Canada, Residential Indoor Air Quality Online
J. Heland, J. Kleffmann, R. Kurtenbach and P. Wiesen,		Guideline: Nitrogen Dioxide, DOI: 10.1039/D1AN01089A
Environ. Sci. Technol., 2001, 35 , 3207–3212.		https://www.canada.ca/en/health-
B. J. Finlayson-Pitts, L. M. Wingen, A. L. Sumner, D. Syomin		canada/services/publications/healthy-living/residential-
and K. A. Ramazan, Phys. Chem. Chem. Phys., 2003, 5, 223-		indoor-air-quality-guideline-nitrogen-dioxide.html.
242.	34	E. Gómez Alvarez, M. Sörgel, S. Gligorovski, S. Bassil, V.
W. Yang, C. Han, H. Yang and X. Xue, Environ. Pollut., 2018,		Bartolomei, B. Coulomb, C. Zetzsch and H. Wortham,
243 , 679–686.		Atmos. Environ., 2014, 95 , 391–399.
A. F. Khalizov, M. Cruz-Quinones and R. Zhang, J. Phys.	35	C. W. Spicer, R. W. Coutant, G. F. Ward, D. W. Joseph, A. J.
Chem. A, 2010, 114 , 7516–7524.		Gaynor and I. H. Billick, Environ. Int., 1989, 15, 643–654.
R. Kurtenbach, K. H. Becker, J. A. G. Gomes, J. Kleffmann, J.	36	T. Wainman, C. J. Weschler, P. J. Lioy and J. Zhang, Environ.
C. Lörzer, M. Spittler, P. Wiesen, R. Ackermann, A. Geyer		Sci. Technol., 2001, 35 , 2200–2206.
and U. Platt, Atmos. Environ., 2001, 35 , 3385–3394.	37	J. Kleffmann, K. H. Becker and P. Wiesen, Atmos. Environ.,
J. G. Calvert, G. Yarwood and A. M. Dunker, Res. Chem.		1998, 32 , 2721–2729.
Intermed., 1994, 20 , 463–502.	38	J. N. Pitts, D. Grosjean, K. Van Cauwenberghe, J. P. Schmid
Q. Yang, H. Su, X. Li, Y. Cheng, K. Lu, P. Cheng, J. Gu, S. Guo,		and D. R. Fitz, <i>Environ. Sci. Technol.</i> , 1978, 12 , 946–953.
M. Hu, L. Zeng, T. Zhu and Y. Zhang, Sci. China Chem., 2014,	39	M. Sleiman, L. A. Gundel, J. F. Pankow, P. Jacob, B. C. Singer
57 , 1032–1042.		and H. Destaillats, Proc. Natl. Acad. Sci. U. S. A., 2010, 107,
Z. Xu, T. Wang, J. Wu, L. Xue, J. Chan, Q. Zha, S. Zhou, P. K.		6576–6581.
K. Louie and C. W. Y. Luk, Atmos. Environ., 2015, 106, 100-	40	T. C. Vandenboer, C. J. Young, R. K. Talukdar, M. Z.
109.		Markovic, S. S. Brown, J. M. Roberts and J. G. Murphy, Nat.
V. Selimovic, R. J. Yokelson, C. Warneke, J. M. Roberts, J. A.		Geosci., 2015, 8 , 55–60.
de Gouw and D. W. T. Griffith, Atmos. Chem. Phys., 2018,	41	G. Huang, X. Zhou, G. Deng, H. Qiao and K. Civerolo, Atmos.
18 , 2929–2948.		Environ., 2002, 36 , 2225–2235.
R. M. Mushinski, R. P. Phillips, Z. C. Payne, R. B. Abney, I. Jo,	42	C. Afif, C. Jambert, V. Michoud, A. Colomb, G. Eyglunent, A.
S. Fei, S. E. Pusede, J. R. White, D. B. Rusch and J. D. Raff,		Borbon, V. Daële, J. F. Doussin and P. Perros, J. Environ. Sci.
Proc. Natl. Acad. Sci., 2019, 116 , 2138–2145.		(China), 2016, 40 , 105–113.
P. Veres, J. M. Roberts, I. R. Burling, C. Warneke, J. De	43	U. Platt, D. Perner, G. W. Harris, A. M. Winer and J. N. Pitts,
Gouw and R. J. Yokelson, J. Geophys. Res. Atmos., 2010,		Nature, 1980, 285 , 312–314.
115.	44	A. A. Salem, A. A. Soliman and I. A. El-Haty, Air Qual.
R. Oswald, T. Behrendt, M. Ermel, D. Wu, H. Su, Y. Cheng,		Atmos. Heal., 2009, 2 , 133–145.
C. Breuninger, A. Moravek, E. Mougin, C. Delon, B. Loubet,	45	T. Górecki and J. Namienik, TrAC - Trends Anal. Chem.,
A. Pommerening-Röser, M. Sörgel, U. Pöschl, T. Hoffmann,		2002, 21 , 276–291.
M. O. Andreae, F. X. Meixner and I. Trebs, Science (80).,	46	F. Wania and C. Shunthirasingham, Environ. Sci. Process.
2013, 341 , 1233–1235.		Impacts, 2020, 22 , 1925–2002.
L. Kramer, L. Crilley, T. Adams, S. Ball, F. Pope and W. Bloss,	47	R. M. Cox, Environ. Pollut., 2003, 126 , 301–311.
Atmos. Chem. Phys. Discuss., 2019, 1–31.	48	M. R. Heal, D. P. H. Laxen and B. B. Marner, Atmosphere
D. B. Collins, R. F. Hems, S. Zhou, C. Wang, E. Grignon, M.		(Basel)., 2019, 10 , 1–28.
Alavy, J. A. Siegel and J. P. D. Abbatt, Environ. Sci. Technol.,	49	A. Loudet and K. Burgess, <i>Chem. Rev.</i> , 2007, 107 , 4891–
2018, 52 , 12419–12427.		4932.
K. Lee, J. Xue, A. S. Geyh, H. Özkaynak, B. P. Leaderer, C. J.	50	T. Ehrenschwender and H. Wagenknecht, 1–13.
Weschler and J. D. Spengler, Environ. Health Perspect.,	51	N. Dorh, S. Zhu, K. B. Dhungana, R. Pati, F. T. Luo, H. Liu and
2002, 110 , 145–149.		A. Tiwari, <i>Sci. Rep.</i> , 2015, 5 , 1–10.
B. P. Leaderer, L. Naeher, T. Jankun, K. Balenger, T. R.	52	N. Boens, V. Leen and W. Dehaen, Chem. Soc. Rev., 2012,
Holford, C. Toth, J. Sullivan, J. M. Wolfson and P. Koutrakis,		41 , 1130–1172.
Environ. Health Perspect., 1999, 107 , 223–231.	53	L. Li, J. Han, B. Nguyen and K. Burgess, J. Org. Chem., 2008,
A. Febo and C. Perrino, Atmos. Environ. Part A, Gen. Top.,		73 , 1963–1970.
1991, 25 , 1055–1061.	54	A. Treibs and FH Kreuzer, Justus Liebigs Ann. Chem.,
T. C. VandenBoer, S. S. Brown, J. G. Murphy, W. C. Keene,		1968, 718, 208–223.
C. J. Young, A. A. P. Pszenny, S. Kim, C. Warneke, J. A. de	55	B. Gu, L. Huang, J. Hu, J. Liu, W. Su, X. Duan, H. Li and S.
Gouw, J. R. Maben, N. L. Wagner, T. P. Riedel, J. A.		Yao, <i>Talanta</i> , 2016, 152 , 155–161.
Thornton, D. E. Wolfe, W. P. Dubé, F. Oztürk, C. A. Brock, N.	56	N. Adarsh, M. Shanmugasundaram and D. Ramaiah, Anal.
Grossberg, B. Lefer, B. Lerner, A. M. Middlebrook and J. M.	_	Chem., 2013, 85 , 10008–10012.
Roberts, J. Geophys. Res. Atmos., 2013, 118 , 10,155-	57	K. J. Huang, W. Z. Xie, H. S. Zhang and H. Wang, <i>Microchim</i> .
10,1/1.		Acta, 2008, 161 , 201–207.
K. W. Wong, B. L. Leter, B. Rappenglück and J. Stutz, Atmos.	58	J. Zhang, F. Pan, Y. Jin, N. Wang, J. He, W. Zhang and W.
cnem. Phys., 2011, 11 , 3595–3609.		Znao, Dye. Pigment., 2018, 155 , 276–283.

Page 13 of 13

 Analyst

D	2	5	0	10
r	a	μ	e	ι.

View Article Online DOI: 10.1039/D1AN01089A

H. Zollinger, Diazo Chemistry I Diazo Chemistry I: Aromatic

This journal is © The Royal Society of Chemistry 20xx